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# The Arctic picoeukaryote *Micromonas pusilla* benefits synergistically from warming and ocean acidification

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**Abstract.** In the Arctic Ocean, climate change effects such as warming and ocean acidification (OA) are manifesting faster than in other regions. Yet, we are lacking a mechanistic understanding of the interactive effects of these drivers on Arctic primary producers. In the current study, one of the most abundant species of the Arctic Ocean, the prasinophyte Micromonas pusilla, was exposed to a range of different pCO<sub>2</sub> levels at two temperatures representing realistic current and future scenarios for nutrient-replete conditions. We observed that warming and OA synergistically increased growth rates at intermediate to high pCO<sub>2</sub> levels. Furthermore, elevated temperatures shifted the pCO<sub>2</sub> optimum of biomass production to higher levels. Based on changes in cellular composition and photophysiology, we hypothesise that the observed synergies can be explained by beneficial effects of warming on carbon fixation in combination with facilitated carbon acquisition under OA. Our findings help to understand the higher abundances of picoeukaryotes such as M. pusilla

#### 1 Introduction

With the progress in using molecular tools to describe marine biodiversity in recent decades, the scientific community has become increasingly aware of the underestimated importance of picoeukaryotes, for both primary and export production of the world's oceans (Richardson and Jackson, 2007; Worden and Not, 2008). Larger phytoplankton such as diatoms are efficient vectors for carbon export due to aggregate formation

under OA, as has been observed in many mesocosm studies.

and ingestion by large zooplankton, leading to the production of fast-settling faecal pellets (Sherr et al., 2003). In contrast, picoeukaryotes are mainly grazed by smaller heterotrophic protists such as ciliates, which have a low carbon retention, excrete relatively more dissolved material and thus fuel recycled production (Sherr and Sherr, 2002). Hence, changes in the relative abundance of pico- and nanoeukaryotes can have large implications for food webs and biogeochemistry (Worden et al., 2015).

Picoeukaryotes tend to dominate low nutrient environments, which is often attributed to their high surface: volume ratios and mixotrophic capacities (Raven, 1998; McKie-Krisberg and Sanders, 2014). The low nutrient concentrations in the Arctic surface ocean, for example, cause picoeukaryotes to be particularly successful in this region. In fact, the globally occurring prasinophyte *Micromonas pusilla* is considered the most abundant species in the Arctic ocean (Šlapeta et al., 2006; Lovejoy et al., 2007; Marquardt et al., 2016). In this environment, strong stratification causes low nutrient concentrations throughout the summer and autumn months (Tremblay et al., 2015), and the occurrence of the polar night requires organisms to either form resting stages or to have heterotrophic capacities (Tremblay et al., 2009; Lovejoy, 2014; Berge et al., 2015; Vader et al., 2015).

Climate change effects manifest faster in the Arctic than anywhere else on the planet (Stocker, 2014). In this region, for example, temperatures are rising more than twice as fast as throughout the rest of the globe (Miller et al., 2010). The concurrent rapid reduction in ice cover allows for more light penetration and longer growing seasons, while

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increased stratification due to ice melt and warming constrain nutrient supply to surface waters, both of which will change the dynamics of primary production (Arrigo et al., 2008; Wassmann and Reigstad, 2011). Ocean acidification (OA) is also especially pronounced in the Arctic Ocean, because low temperatures and alkalinity make the system sensitive to anthropogenic CO<sub>2</sub> loading (AMAP, 2013; Qi et al., 2017). Picoeukaryotes such as M. pusilla may benefit from these changes and are considered potential winners of climate change. In the Canadian Arctic, for example, picoeukaryote abundances are increasing as surface waters get warmer, fresher and more oligotrophic (Li et al., 2009). Regarding OA effects, the majority of studies on natural phytoplankton assemblages have shown that picoeukaryotes, particularly M. pusilla, increase in relative abundance with increasing pCO<sub>2</sub> levels (Engel et al., 2008; Meakin and Wyman, 2011; Newbold et al., 2012; Brussaard et al., 2013; Hussherr et al., 2017; Schulz et al., 2017). Despite the evident sensitivity of M. pusilla to changes in pCO<sub>2</sub> levels, a detailed assessment of the OA effects, their interaction with warming as well as the underlying mechanisms in this important species is still missing.

Like all photosynthetic organisms, cells of M. pusilla need to maintain a balance between energy sources (i.e. light harvesting by the photosynthetic apparatus) and sinks (most importantly carbon fixation in the Calvin cycle) to prevent harmful levels of excitation pressure on the photosynthetic electron transport chain (Behrenfeld et al., 2008). Light harvesting and electron transport in the photosystems are largely independent of changes in temperature and  $pCO_2$  (Mock and Hoch, 2005; Hoppe et al., 2015), but the impact of these drivers on energy sinks can potentially affect the energy balance of the cell: the beneficial effects of elevated pCO<sub>2</sub> observed in phytoplankton are thought to be caused by increased diffusive CO2 supply, reduced CO2 leakage or lowered costs to operate their CO<sub>2</sub> concentrating mechanisms (Rost et al., 2008; Bach et al., 2013). Elevated temperatures, however, can change enzyme kinetics, including those involved in the Calvin cycle, thus leading to a larger sink of excitation energy (Maxwell et al., 1994; Toseland et al., 2013). Hence, both ocean warming and acidification potentially increase the efficiency of photosynthesis and biomass production, at least up to the organisms' respective optimum levels. Above these levels, temperatures and proton concentrations start to disrupt enzymatic processes, increase the investment into pH homeostasis, and impair the delicate regulation of cellular processes (Levitt, 1980; Taylor et al., 2001; Flynn et al., 2012). Thus, the complex balance between beneficial and detrimental effects will determine whether the combination of warming and OA will synergistically promote or deteriorate phytoplankton growth and biomass build-up.

In the current study, we aim to investigate the responses of an Arctic M. pusilla strain to warming and OA. To this end, M. pusilla was grown at four  $pCO_2$  levels ranging from preindustrial to future scenarios (180–1400  $\mu$ atm) under 2 and

6°C, which represent the magnitude of the projected future temperature increase in this region (Collins et al., 2013), but also the current spring and summer temperatures in the environment where the strain was isolated (Hegseth et al., 2018).

#### 2 Material & methods

### 2.1 Culture conditions

Monoclonal cultures of the picoeukaryote Micromonas pusilla (Butcher) Manton & Parke (isolated in 2014 by Klara Wolf in Kongsfjorden, Svalbard, 79° N; local temperature range -1.5 to 8 °C) taxonomic identification confirmed by rDNA sequencing of SSU, LSU and ITS sequences) were grown in 1 L glass bottles in semi-continuous dilute-batch cultures (maximum 129 000 cells mL<sup>-1</sup>; diluted every 3-4 days) under constant irradiances of  $150 \pm 26 \,\mu\text{mol photons m}^{-2}\,\text{s}^{-1}$ . Media consisted of  $0.2 \,\mu\text{m}$ sterile-filtered Arctic seawater with a salinity of 32.7 enriched with macronutrients, trace metals and vitamins according to  $F/2_R$  medium (Guillard and Ryther, 1962). Light intensities were provided by daylight lamps (Philips Master TL-D 18 W; emission peaks at wavelengths of 440, 560 and 635 nm), adjusted by neutral density screens and monitored using a LI-1400 data logger (Li-Cor) equipped with a  $4\pi$  sensor (Walz). Cells were growing at four different CO<sub>2</sub> partial pressures (pCO<sub>2</sub>; 180, 380, 1000 and 1400 µatm) and two temperatures ( $2.2 \pm 0.3$  and  $6.3 \pm 0.2$  °C). Cultures were acclimated to these conditions for at least seven generations prior to sampling.

Different  $p\text{CO}_2$  conditions were achieved by aeration of the incubation bottles with air of the respective  $p\text{CO}_2$  levels delivered through sterile 0.2 µm air filters (Midisart 2000, Sartorius Stedim) for 24 h prior to inoculation. Gas mixtures were generated using a gas flow controller (CGM 2000 MCZ Umwelttechnik), in which CO<sub>2</sub>-free air (<1 ppmv CO<sub>2</sub>; Dominick Hunter) was mixed with pure CO<sub>2</sub> (Air Liquide Deutschland). The  $p\text{CO}_2$  levels in the gas mixtures were regularly monitored with a non-dispersive infrared analyzer system (LI6252, LI-COR Biosciences), calibrated with CO<sub>2</sub>-free air and purchased gas mixtures of  $150 \pm 10$  and  $1000 \pm 20$  ppmv CO<sub>2</sub> (Air Liquide Deutschland).

#### 2.2 Carbonate chemistry

Samples for total alkalinity ( $A_{\rm T}$ ) were filtered through 0.7 µm glass fibre filters (GF/F, Whatman) and stored in borosilicate bottles at 3 °C.  $A_{\rm T}$  was estimated from duplicate potentiometric titration (Brewer et al., 1986) using a TitroLine alpha plus (Schott Instruments).  $A_{\rm T}$  values were corrected for systematic errors based on measurements of certified reference materials (CRMs provided by Andrew Dickson, Scripps, USA; batch no. 111; reproducibility  $\pm$  5 µmol kg $^{-1}$ ). Total dissolved inorganic carbon ( $C_{\rm T}$ ) samples were filtered through 0.2 µm cellulose-acetate filters (Sartorius Ste-

dim) and stored in gas-tight borosilicate bottles at 3 °C. C<sub>T</sub> was measured colorimetrically in triplicates with a QuAAtro autoanalyzer (Seal; Stoll et al., 2001). The analyser was calibrated with NaHCO<sub>3</sub> solutions (with a salinity of 35, achieved by addition of NaCl) to achieve concentrations ranging from 1800 to 2300  $\mu$ mol  $C_{\rm T}$  kg<sup>-1</sup>. CRMs were used for corrections of errors in instrument performance such as baseline drifts (reproducibility  $\pm 8 \,\mu \text{mol kg}^{-1}$ ). Seawater pH<sub>total</sub> was measured potentiometrically with a two-point calibrated glass reference electrode (IOline, Schott Instruments). An internal TRIS-based reference standard (Dickson et al., 2007) was used to correct for variability on electrode performance (reproducibility  $\pm 0.015$  pH units). Following recommendations by Hoppe et al. (2012), seawater carbonate chemistry including  $pCO_2$  was calculated from  $A_T$  and pH using CO<sub>2SYS</sub> (Pierrot et al., 2006). The dissociation constants of carbonic acid of Mehrbach et al. (1973), as refitted by Dickson and Millero (1987), were used for calculations. Dissociation constants for KHSO<sub>4</sub> were taken from Dickson (1990).

# 2.3 Growth, elemental composition and production rates

Samples for cell counts were fixed with glutaraldehyde (0.5% final concentration). After gentle mixing, samples were stored at room temperature in the dark for 15 min, and subsequently frozen in liquid nitrogen and stored at -80 °C. Prior to analysis, samples were thawed on ice and mixed thoroughly. After addition of 10 µL SybrGreen working solution (dissolved in DMSO) and 10 µL YG beads working solution (1 µm Fluoresbrite calibration beads grade YG, Polyscience), samples were counted on an Accuri C6 flow cytometer (BD Biosciences) equipped with a blue solid-state laser (488 nm excitation wavelength) run on medium fluidics settings (35 µL min<sup>-1</sup>; 16 µm core size) with a limit of 50 000 events or 250 µL. Analysis was performed based on red (FL3) channel,  $> 670 \,\mathrm{nm}$ ) and green (FL1 channel,  $533 \pm 30 \,\mathrm{nm}$ ) fluorescence, as well as sideward and forward light scattering. Specific growth rates constants ( $\mu$ ) were determined from exponential fits of cell counts over 4 consecutive days.

Particulate organic carbon (POC) and nitrogen (PON) were measured after filtration onto precombusted (15 h,  $500\,^{\circ}$ C) GF/F filters (Whatman), stored at  $-20\,^{\circ}$ C and dried for at least 12 h at  $60\,^{\circ}$ C prior to sample preparation. Analysis was performed using a CHNS-O elemental analyser (Euro EA 3000, HEKAtech). Contents of POC and PON were corrected for blank measurements and normalised to filtered volume and cell densities to yield cellular quotas. Production rates of POC were calculated by multiplying the cellular quota with the division rate constant k of the respective incubation. Samples for determination of chlorophyll a (Chl a) were filtered onto GF/F filters (Whatman), immediately placed into liquid nitrogen and stored at  $-80\,^{\circ}$ C until analysis. Chl a was subsequently extracted in  $8\,^{\circ}$  mL  $90\,^{\circ}$ 

acetone at 4 °C over night. Chl *a* concentrations were determined on a fluorometer (TD-700, Turner Designs), using an acidification step (1 M HCl) to determine phaeopigments (Knap et al., 1996).

### 2.4 Variable Chl a fluorescence

Photophysiological characteristics, based on photosystem II (PSII) variable Chl a fluorescence, were measured using a fast repetition rate fluorometer (FRRf; FastOcean PTX, Chelsea Technologies) in combination with a FastAct Laboratory system (Chelsea Technologies). The excitation wavelength of the fluorometer's light-emitting diodes (LEDs) was 450 nm, and the applied light intensity was 21 587  $\mu$ mol photons m<sup>-2</sup> s<sup>-1</sup>. The FRRf was used in single turnover mode, with a saturation phase comprising 100 flashlets on a 2  $\mu$ s pitch and a relaxation phase comprising 40 flashlets on a 50  $\mu$ s pitch. Measurements from all replicates (n=3) were conducted in a temperature-controlled chamber ( $\pm$  0.2 °C) at the respective treatment temperature.

After subtraction of a blank value, the minimum ( $F_0$  and  $F_0'$  for light- and dark-acclimated measurements, respectively) and maximum Chl a fluorescence ( $F_{\rm m}$ , and  $F_{\rm m}'$  for light- and dark-acclimated measurements, respectively) were estimated from iterative algorithms for the induction (Kolber et al., 1998) and relaxation phase (Oxborough, 2012) after 15 min of dark acclimation, which was sufficient to achieve a dark-acclimated state (data not shown). All fluorescence parameters were calculated by standard equations (Genty et al., 1989; Maxwell and Johnson, 2000). Maximum quantum yields of PSII (apparent PSII photochemical quantum efficiency;  $F_{\rm v}/F_{\rm m}$ ) were calculated as

$$F_{\rm v}/F_{\rm m} = (F_{\rm m} - F_0)/F_{\rm m}.$$
 (1)

Fluorescence-based photosynthesis–irradiance curves (PI) were conducted at six irradiances (I) between 33 and 672 µmol photons m<sup>-2</sup> s<sup>-1</sup>, with an acclimation time of 10 min per light step. Electron transfer rate (ETR, measured in mol e<sup>-</sup> (mol RCII)<sup>-1</sup> s<sup>-1</sup>) through PSII for each light step was calculated as follows:

$$ETR = ((F'_{m} - F'_{0})/F'_{m}) \cdot I.$$
(2)

Following the suggestion Silsbe bv and Kromkamp (2012),the light-use efficiency  $(\alpha,$  $mole^-m^2 (molRCII)^{-1} (molphotons)^{-1})$ the maximum electron transfer rates per RCII (ETR<sub>max</sub>,  $mole^{-}$  (molRCII)<sup>-1</sup> s<sup>-1</sup>) were estimated by fitting the data to the model by Webb et al. (1974):

$$ETR = ETR_{max} \cdot [1 - e^{(-(\alpha \cdot I)/ETR_{max})}]. \tag{3}$$

The light saturation index  $(E_k, \mu \text{mol photons m}^{-2} \text{s}^{-1})$  was then calculated as  $\text{ETR}_{\text{max}}/\alpha$ . Maximum non-photochemical quenching (NPQ) of Chl a fluorescence at

**Table 1.** Seawater carbonate chemistry at the end of the experiments (n = 3; mean  $\pm 1$  SD). CO<sub>2</sub> partial pressure (pCO<sub>2</sub>) and dissolved CO<sub>2</sub> concentrations were calculated from total alkalinity ( $A_T$ ) and pH<sub>total</sub> at 2 or 6 °C, a salinity of 32.7 using CO<sub>2</sub>SYS (Pierrot et al., 2006), and phosphate and silicate concentrations of 10 and 100  $\mu$ mol kg<sup>-1</sup>, respectively. NA indicates that values are not available for this specific treatment.

Temperature (°C)	pCO <sub>2</sub> level (µatm)	pH total scale	$A_{\rm T}$ (µmol kg <sup>-1</sup> )	$C_{\rm T}$ (µmol kg <sup>-1</sup> )	dissolved $CO_2$ ( $\mu mol  kg^{-1}$ )	pCO <sub>2</sub> (μatm)
2	180 380 1000 1400	$8.3 \pm 0.01$ $8.11 \pm 0.01$ $7.68 \pm 0.01$ $7.52 \pm 0.02$	$2264 \pm 9  2244 \pm 30  2255 \pm 45  2243 \pm 5$	2024 ± 6 2124 ± 11 2215 ± 23 NA	$11.6 \pm 0.2$ $19.0 \pm 0.7$ $56.4 \pm 1.3$ $81.1 \pm 3.1$	$   \begin{array}{c}     197 \pm 3 \\     323 \pm 12 \\     959 \pm 22 \\     1380 \pm 53   \end{array} $
6	180 380 1000 1400	$8.3 \pm 0.01$ $8.04 \pm 0.01$ $7.65 \pm 0.01$ $7.52 \pm 0.01$	$2243 \pm 28  2256 \pm 21  2262 \pm 22  2265 \pm 5$	$1969 \pm 10$ $2058 \pm 7$ $2178 \pm 14$ NA	$10.0 \pm 0.3$ $20.0 \pm 0.5$ $52.6 \pm 1.6$ $73.6 \pm 0.9$	$   \begin{array}{c}     198 \pm 6 \\     394 \pm 10 \\     1036 \pm 31 \\     1449 \pm 18   \end{array} $

**Table 2.** Growth rate constants  $\mu$ , division rate constants k, POC production rates and cellular quota of Chl a, and POC and PON as well as their ratios of M. pusilla at the end of the experiment under the different treatment conditions (n = 3; mean  $\pm 1$  SD). Results from statistical analysis can be found in Table S2.

Temperature	pCO <sub>2</sub>	Growth rate constant $\mu$	Division rate constant k	POC production	POC quota	PON quota	Chl a quota	POC:PON	POC : Chl a
(°C)	(µatm)	$(day^{-1})$	$(day^{-1})$	$(fmolcell^{-1}day^{-1})$	$(fmol cell^{-1})$	$(fmol cell^{-1})$	$(fg cell^{-1})$	$(molmol^{-1})$	$(gg^{-1})$
	180	$0.75 \pm 0.04$	$1.08 \pm 0.05$	$256 \pm 11$	$239 \pm 20$	$28.7 \pm 2.6$	$28.6 \pm 2.1$	$8.3 \pm 0.1$	100 ± 7
2	380	$0.85 \pm 0.03$	$1.23 \pm 0.04$	$290 \pm 16$	$237 \pm 22$	$30.9 \pm 1.8$	$24.7 \pm 1.4$	$7.7 \pm 0.3$	$115 \pm 10$
2	1000	$0.79 \pm 0.05$	$1.15 \pm 0.07$	$224 \pm 23$	$196 \pm 25$	$24.9 \pm 4.9$	$20.0 \pm 2.9$	$8.0 \pm 0.6$	$118 \pm 5$
	1400	$0.82 \pm 0.05$	$1.18\pm0.07$	$235 \pm 13$	$199 \pm 16$	$23.9 \pm 2.3$	$22.0\pm1.5$	$8.4 \pm 0.1$	$109 \pm 4$
6	180	$1.06 \pm 0.03$	$1.53 \pm 0.05$	$376 \pm 15$	$245 \pm 3$	$26.9 \pm 0.6$	$21.1 \pm 0.8$	$9.1 \pm 0.1$	140 ± 7
	380	$1.05 \pm 0.03$	$1.52 \pm 0.04$	$342 \pm 39$	$226 \pm 26$	$25.9 \pm 2.5$	$22.1 \pm 2.6$	$8.7 \pm 0.3$	$123 \pm 11$
	1000	$1.25 \pm 0.05$	$1.80 \pm 0.07$	$497 \pm 12$	$275 \pm 6$	$33.7 \pm 2.0$	$27.2 \pm 0.9$	$8.2 \pm 0.6$	$122 \pm 6$
	1400	$1.05 \pm 0.04$	$1.52 \pm 0.06$	$400 \pm 14$	$263 \pm 7$	$31.1\pm1.7$	$28.6 \pm 3.2$	$8.5 \pm 0.3$	$111\pm16$

irradiances of 672 µmol photons m<sup>-2</sup> s<sup>-1</sup> (i.e. the highest irradiance step of the PI curve) were calculated using the normalised Stern–Volmer coefficient, also termed NSV, as described in McKew et al. (2013):

$$(F_a'/F_v') - 1 = F_0'/F_v', (4)$$

where  $F'_q$  is the differences between measured and maximal fluorescence (Suggett et al., 2010).  $F'_0$  was measured after each light step (with a duration of 90 s).

### 2.5 Statistics

All data are given as the mean of three biological replicates with  $\pm 1$  standard deviation. To test for significant differences between the treatments, two-way analyses of variance (ANOVA) with additional normality (Kolmogorov–Smirnov) and post hoc (Holm–Šidák) tests were performed. The significance level was set to 0.05. Statistical analyses were performed with the program SigmaPlot (Systat Software Inc, Version 12.5).

#### 3 Results

# 3.1 Carbonate chemistry

Regular dilution of cultures with a pre-aerated seawater medium kept carbonate chemistry stable over the course of the experiment. More specifically, in each bottle the drift in  $A_{\rm T}$  and  $C_{\rm T}$  compared to initial values was  $\leq 3$  and  $\leq 4\%$ , respectively (data not shown). Final carbonate chemistry in the 2 °C treatments yielded pCO<sub>2</sub> levels of  $197 \pm 3$ ,  $323 \pm 12$ ,  $959 \pm 22$  and  $1380 \pm 53 \,\mu atm$  (Table 1). In the 6 °C treatments,  $pCO_2$  levels were  $198 \pm 6$ ,  $394 \pm 10$ ,  $1036 \pm 31$  and  $1449 \pm 18 \,\mu atm$ . Please note that the same  $pCO_2$  level translates into differing dissolved CO<sub>2</sub> concentrations at different temperatures due to the temperature dependency of the carbonate system. Specifically, the treatment pCO2 values translated into up to 13 % lower dissolved CO<sub>2</sub> concentrations in the 6°C compared to the 2°C treatment (Table 1; see Fig. S1 in the Supplement). Concurrently, the  $pCO_2$  levels at 2 °C corresponded to pH<sub>total</sub> values of  $8.30 \pm 0.01$ ,  $8.11 \pm 0.01$ ,  $7.68 \pm 0.01$  and  $7.52 \pm 0.02$ , respectively. In the 6°C treatment, pH<sub>total</sub> values of the four pCO<sub>2</sub> treatments

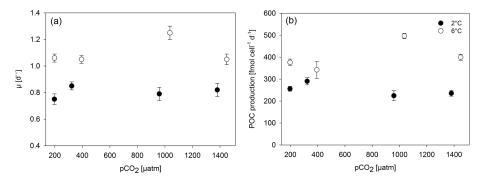


Figure 1. Specific growth rate constant  $\mu$  (a) and POC production (b) of *M. pusilla* under low (filled symbols) and high (open symbols) temperatures as a function of  $pCO_2$  (n=3; mean  $\pm 1$  SD). Results from statistical analysis can be found in Table S2.

were  $8.30 \pm 0.01$ ,  $8.04 \pm 0.01$ ,  $7.65 \pm 0.01$  and  $7.52 \pm 0.01$ , respectively.

#### 3.2 Growth and biomass build-up

Growth rates constants of exponentially growing M. pusilla cultures were significantly affected by the applied treatments (Fig. 1, Tables 2, S1 in the Supplement). Depending on the pCO<sub>2</sub> level, temperature increased growth by 20 to 60 % with an average of  $0.80\,\mathrm{day}^{-1}$  under low and  $1.10\,\mathrm{day}^{-1}$ under high temperature conditions (two-way ANOVA, F =328, p < 0.001). Overall, there was also a positive  $pCO_2$  effect on growth (two-way ANOVA, F = 9, p = 0.001), even though no linear trends with either  $pCO_2$  or  $[CO_2]$  were observed (Figs. 1, S1). The observed pCO<sub>2</sub> responses also differed between temperature levels, indicating a significant interaction between both drivers (two-way ANOVA, F = 12, p < 0.001): under low temperature, growth increased significantly from 180 to 380  $\mu$ atm  $pCO_2$  (post hoc, t = 3.1, p = 0.04), while there was a declining yet insignificant trend in growth with further increases in pCO<sub>2</sub>. Under high temperature, growth was significantly higher under 1000 compared to lower (180  $\mu$ atm; post hoc, t = 5.6, p < 0.001) and higher pCO<sub>2</sub> levels (1400  $\mu$ atm; post hoc, t = 5.9, p <0.001). Thus, warming shifted the optimum  $pCO_2$  range for growth to higher levels (Fig. 1a).

This trend was also observed in terms of POC production rates (Fig. 1b, Table 2), with significant effects of temperature (Table S1; two-way ANOVA, F=356, p<0.001),  $p\text{CO}_2$  (two-way ANOVA, F=7, p=0.003), and their interaction (two-way ANOVA, F=29, p<0.001). At low temperatures, higher production rates were observed at 180 and 380 µatm compared to those at 1000 and 1400 µatm  $p\text{CO}_2$  (post hoc tests, t=3.5, p=0.016 and t=3.0, p=0.046, respectively). At high temperatures, POC production rates were significantly higher at 1000 µatm than at all other  $p\text{CO}_2$  levels (post hoc tests, e.g. t=9.1, p<0.001 for 380 vs. 1000 µatm and t=7.4, p<0.001 for 1000 vs. 1400 µatm), again indicating an upward shift in the  $p\text{CO}_2$  optimum with warming.

## 3.3 Cellular composition

Overall, POC quotas (Fig. 2a, Tables 2, S1) were significantly higher under elevated compared to low temperatures (two-way ANOVA, F = 24, p < 0.001), but no overarching trend with  $p\text{CO}_2$  was observed. Under low temperatures, cells had significantly higher POC quotas at low  $p\text{CO}_2$  levels (180 and 380  $\mu$ atm) compared to high  $p\text{CO}_2$  levels (1000 and 1400  $\mu$ atm; all four post hoc tests were significant, e.g. 380 vs. 1000  $\mu$ atm: t = 2.8, p = 0.033). This trend reversed under high temperatures, where POC quota were highest under 1000 and 1400  $\mu$ atm (post hoc test, t = 3.5, p = 0.024). Thus, temperature and  $p\text{CO}_2$  levels exhibited a significant interactive effect on POC quota (two-way ANOVA, F = 10, p < 0.001).

Similar trends were observed in terms of cellular PON quota (Fig. 2b, Tables 2, S1), where temperature (two-way ANOVA, F = 5, p = 0.045) and its interaction with  $pCO_2$  (two-way ANOVA, F = 10, p < 0.001) significantly affected the results. Here, opposing  $pCO_2$  effects under different temperatures were more subtle, with PON quotas under low temperatures only being significantly decreased between 380 and 1400  $\mu$ atm (post hoc test, t = 3.3, p = 0.027), while under high temperatures PON quotas significantly increased from 180 and 380 to 1000  $\mu$ atm  $pCO_2$  (post hoc tests, t = 3.7, p = 0.012 and t = 2.8, p = 0.028, respectively).

Regarding cellular Chl a quota, there were no significant effects of temperature or  $p\text{CO}_2$  alone (Fig. 2c, Tables 2, S1), but a significant interaction between the two drivers was found (two-way ANOVA, F=18, p<0.001): under low temperatures, Chl a quotas decreased from low (180  $\mu$ atm) to high  $p\text{CO}_2$  levels (1000 and 1400  $\mu$ atm; post hoc tests, t=5.0, p<0.001 and t=3.9, p=0.006, respectively). Under high temperatures, the opposite trend was observed, where Chl a quotas increased from low (180 and 380  $\mu$ atm) to high  $p\text{CO}_2$  levels (1000 and 1400  $\mu$ atm; all four post hoc tests significant, e.g. 380 vs. 1000  $\mu$ atm: t=3.0, p=0.027).

Molar C: N ratios of the biomass (Tables 2, S1) increased with temperature (two-way ANOVA, F = 14, p = 0.002), yet this overall difference was mainly driven by results at

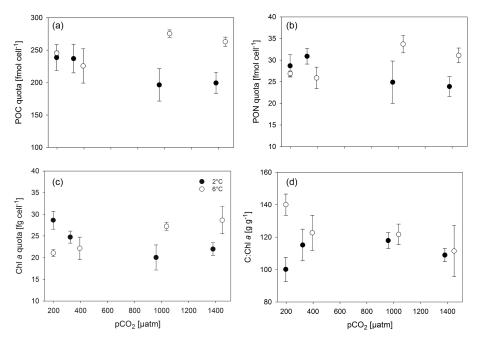


Figure 2. Cellular composition, i.e. POC (a), PON (b) and Chl a quota (c) as well as C: Chl a ratios (d), of M. pusilla under low (filled symbols) and high (open symbols) temperatures as a function of  $pCO_2$  (n = 3; mean  $\pm 1$  SD). Results from statistical analysis can be found in Table S2.

low  $p\text{CO}_2$  levels (180 and 380  $\mu$ atm; post hoc tests, t = 2.7, p = 0.017 and t = 3.5, p = 0.003, respectively). By itself,  $p\text{CO}_2$  did not significantly affect C: N ratios.

The ratios of C:Chl a (Fig. 2d, Tables 2, S1) were elevated under high compared to low temperature conditions (two-way ANOVA, F=14, p=0.002), an effect that was most pronounced at  $pCO_2$  levels of  $180 \, \mu atm$  (post hoc test, t=5.5, p<0.001). While there was no effect of  $pCO_2$  on C:Chl a at low temperature, C:Chl a decreased with increasing  $pCO_2$  at high temperature (two-way ANOVA, interaction term, F=6, p=0.007;  $180 \, vs.$   $1400 \, \mu atm$  at  $6 \, ^{\circ}C$  post hoc test, t=3.9, p=0.008).

### 3.4 Chl a fluorescence-based photophysiology

The effects of the applied treatments on photophysiology were studied by means of FRRf, which investigates photochemistry at photosystem II (PSII). No effects of the applied treatments were observed in most parameters investigated (Tables 3, S1). This was true for the dark-acclimated quantum yield efficiency of PSII ( $F_{\rm v}/F_{\rm m}$ ), which was similar in all treatments with values of 0.45  $\pm$  0.06, as well as for absorption cross sections of PSII light harvesting ( $\sigma_{\rm PSII}$ ).

Furthermore, the fitted parameters of FRRf-based PI curves ( $\alpha$ , ETR<sub>max</sub> and  $E_K$ ) were independent of the experimental treatments (Tables 3, S1). In contrast, the rate constant of the reopening of PSII reaction centres ( $\tau_{ES}$ ; Tables 3, S1) was slightly yet significantly smaller under high temperatures (two-way ANOVA, F=6, p=0.029), even though

this overall response also depended on the applied  $pCO_2$  levels (two-way ANOVA, interaction term, F = 4, p = 0.033).

Maximum non-photochemical quenching (NPQ<sub>max</sub>; Tables 3, S1) increased significantly with  $pCO_2$  (Table S1; two-way ANOVA, F=0, p=0.002) while temperature had no effect. Post hoc tests revealed that this response was mainly driven by high NPQ<sub>max</sub> values at  $1000 \, \mu atm$ , which were significantly higher than in any other  $pCO_2$  treatment (e.g. t=4.1, p=0.006 for  $380 \, vs. 1000 \, \mu atm$  and t=3.1, p=0.030 for  $1000 \, vs. 1400 \, \mu atm$ ).

# 4 Discussion

#### 4.1 Micromonas pusilla benefits from warming

We observed a strong stimulation of growth rates and biomass build-up with increasing temperature (Fig. 1, Table 2). Even though the isolate stems from 1.8 °C water temperature, the beneficial effects of warming from 2 to 6 °C are not surprising: *M. pusilla* is known to dominate Arctic phytoplankton assemblages in the summer and autumn situations (Lovejoy et al., 2007; Marquardt et al., 2016) when surface temperatures of 6 °C or more can be reached (Hegseth et al., 2018). Our results are also in line with mesocosm experiments that indicate stimulatory effects of warming on picoplankton abundances (Daufresne et al., 2009; Sommer et al., 2015) as well as with the temperature optimum of 6–8 °C observed for another Arctic strain of *M. pusilla* (Lovejoy et al., 2007).

Table 3. FRR fluorometrical PSII photochemistry measurements – PSII quantum yield efficiency  $F_{\rm V}/F_{\rm m}$  (dimensionless), functional absorption cross section ( $\sigma_{\rm PSII}$ ) (nm<sup>-2</sup> PSII<sup>-1</sup>), rate of PSII re-opening ( $\tau_{\rm ES}$ , ms), maximum non-photochemical quenching at 672 μmol photons m<sup>-2</sup> s<sup>-1</sup> (NPQ<sub>max</sub>, dimensionless), maximum light-use efficiency (initial slope α, mol e<sup>-</sup> m<sup>2</sup> (mol RCII)<sup>-1</sup> (mol photons)<sup>-1</sup>), maximal absolute electron transfer rates through PSII (ETR<sub>max</sub>, mol e<sup>-</sup> (mol RCII)<sup>-1</sup> s<sup>-1</sup>), and the light saturation index ( $E_K$ , μmol photons m<sup>-2</sup> s<sup>-1</sup>) under the different temperature and pCO<sub>2</sub> treatments (n = 3; mean ± 1 SD). Results from statistical analyses can be found in Table S2.

Temp.	$pCO_2$	$F_{\rm v}/F_{\rm m}$	$\sigma_{ m PSII}$	$ au_{ ext{ES}}$	NPQ <sub>max</sub>	α	ETR <sub>max</sub>	$E_K$
2	180	$0.50 \pm 0.01$	$8.66 \pm 0.35$	$439 \pm 8$	$2.26 \pm 0.18$	$0.42 \pm 0.05$	$33 \pm 2$	$81 \pm 13$
	380	$0.43 \pm 0.09$	$8.93 \pm 0.26$	$425 \pm 4$	$3.51 \pm 0.55$	$0.32 \pm 0.15$	$25 \pm 5$	$91 \pm 44$
	1000	$0.45 \pm 0.08$	$8.55 \pm 0.07$	$448 \pm 1$	$3.96 \pm 0.71$	$0.42 \pm 0.03$	$31 \pm 2$	$75 \pm 10$
	1400	$0.47 \pm 0.10$	$9.06 \pm 0.05$	$422\pm14$	$2.45 \pm 0.44$	$0.43 \pm 0.08$	$31 \pm 7$	$75 \pm 31$
6	180	$0.49 \pm 0.01$	$9.22 \pm 0.22$	412 ± 6	$2.51 \pm 0.37$	$0.49 \pm 0.08$	$28 \pm 10$	59 ± 28
	380	$0.43 \pm 0.12$	$8.83 \pm 0.17$	$427 \pm 6$	$2.83 \pm 0.59$	$0.38 \pm 0.09$	$35 \pm 14$	$90 \pm 17$
	1000	$0.41 \pm 0.07$	$8.91 \pm 0.22$	$422 \pm 11$	$4.94 \pm 1.46$	$0.33 \pm 0.09$	$32 \pm 5$	$100 \pm 21$
	1400	$0.45 \pm 0.04$	$8.71 \pm 0.50$	$428\pm19$	$2.93 \pm 0.50$	$0.38 \pm 0.04$	$40 \pm 5$	$104 \pm 6$

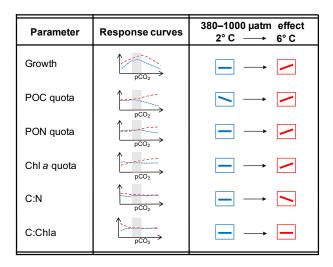
Below the temperature optimum of a cell, warming causes an acceleration of the entire metabolism, as enzymatic reactions run faster under these conditions (Eppley, 1972; Brown et al., 2004). In this study, warming caused higher growth rates, POC quotas and biomass production (Fig. 2, Tables 2, S1), indicating that particularly the fixation and storage of carbon was facilitated by increasing temperature. Electron transport processes, however, were largely independent of temperature (Tables 3, S1). Thus, temperature affected the balance between electron transport ("light reaction") and carbon fixation in the Calvin cycle ("dark reactions"). Especially under relatively low temperatures, as investigated here, warming can decrease the excitation pressure on the electron transport chain of the photosystems by increasing the temperature-limited turnover rates of enzyme reactions such as RuBisCO (Mock and Hoch, 2005). Thus, cells grown under low temperatures need to invest relatively more energy into biosynthesis than into photochemistry compared to cells grown under high temperatures (Toseland et al., 2013). While it has been shown that Antarctic diatoms can compensate for slow RuBisCO kinetics by increasing the expression of this enzyme (Young et al., 2014), it is unknown whether such acclimation responses also occur in prasinophytes. Regarding the C: Chl a ratio, this can be taken as an indicator of how many resources the cell retains as carbon biomass (e.g. structural and storage compounds) relative to how much is invested into its light harvesting capacities (Halsey and Jones, 2015). In this study, the strong temperature-dependent increase in C: Chl a (Fig. 2, Table S1) under potentially limiting  $pCO_2$  levels of 180 µatm suggests that under warming, the balance between light harvesting and carbon fixation was indeed more beneficial for biomass build-up. Furthermore, elevated temperature significantly decreased  $\tau_{ES}$  (Tables 3, S1), which can serve as a proxy of the rate at which downstream processes can remove electrons from PSII (Kolber et al., 1998). Thus, our results indicate that the drainage of

electrons into carbon fixation was faster under warmer conditions, explaining the higher growth and biomass production under these conditions.

# 4.2 Warming shifts CO<sub>2</sub> optima towards higher pCO<sub>2</sub> levels

Under  $6 \,^{\circ}$ C and  $pCO_2$  levels expected to be reached by the end of this century, OA had a significantly positive effect on growth and biomass build-up (Fig. 1). This finding is in line with previous studies, which have shown that picoeukaryotes can benefit strongly from OA in both laboratory and mesocosm studies (Meakin and Wyman, 2011; Newbold et al., 2012; Schaum et al., 2012; Brussaard et al., 2013; Maat et al., 2014; Schulz et al., 2017). Such positive response to OA could indicate that picoeukaryotes such as M. pusilla are mainly dependent on diffusive CO<sub>2</sub> supply and thus directly benefit from higher CO<sub>2</sub> concentrations (Brussaard et al., 2013; Schulz et al., 2013, 2017). While this could be related to the large surface-to-volume ratio of small cells, opposite trends within the group of diatoms (i.e. higher sensitivity of larger compared to smaller diatoms; Wu et al., 2014; Sett et al., 2018) suggest more complex underlying mechanisms at

Despite this overall effect, growth rates of M. pusilla tended to follow a non-linear response curve over the tested range of glacial to elevated future  $pCO_2$  levels (i.e. 180 to 1400  $\mu$ atm), i.e. growth increased with increasing  $pCO_2$  from low to intermediate, but decreased again under the highest  $pCO_2$  levels (Fig. 1). Such an optimum behaviour can be expected for most environmental drivers (Harley et al., 2017) and has previously been observed in response to OA (Sett et al., 2014; Wolf et al., 2018). The response patterns in these studies were attributed to a combination of beneficial effects of rising  $pCO_2$  under potentially carbon-limiting conditions for photosynthesis, and negative effects of declining pH on cellular homeostasis and enzyme performance, which man-



**Figure 3.** Schematic illustration of results for both temperatures over the entire range of  $pCO_2$  levels as well as focusing on the responses between 380 and 1000  $\mu$ atm (as the representation for commonly used OA treatments) and their modulation by temperature.

ifest mainly at high  $pCO_2$  (Bach et al., 2013). This non-linearity in the observed  $pCO_2$  effects emphasises the importance of experiments with more than two  $pCO_2$  levels in order to properly describe OA-response patterns of organisms.

On a more general level, apparent discrepancies between OA studies can be attributed to actual differences in the environmental settings and their interactive effects with pCO<sub>2</sub> (Riebesell and Gattuso, 2015). When comparing the two most commonly applied  $pCO_2$  levels, i.e. the present-day and the anticipated end-of-century situation, the effects of OA on most of the investigated physiological parameters are reversed under 6 compared to 2 °C (Fig. 3). This illustrates how difficult it is to infer responses to OA from experiments applying only one set of environmental conditions. It is also noteworthy that the combination of OA and warming led to more densely packed cells (no change in cell size based on flow cytometric measurements; data not shown) with similar stoichiometry compared to the control treatment (Table 2). This indicates that cells managed to cope well with the experienced future conditions. Furthermore, warming altered the OA-dependent change in most of the investigated parameters in a direction that indicates higher fitness compared to low temperatures (e.g. higher growth rates and higher elemental quota; Fig. 3; Hoppe et al., 2018b). Thus, the increase in growth under future compared to ambient conditions was larger than what would be expected by the respective responses to warming and OA in isolation, indicating synergistic beneficial effects of both drivers.

# **4.3** Potential mechanism underlying the interaction between warming and OA

The observed synergistic effects could be explained by their specific impacts on carbon acquisition and fixation. As outlined in the introduction, light and dark reaction of photosynthesis need to be balanced to achieve high biomass production while avoiding photodamage (Behrenfeld et al., 2008). According to our data, this balance is shifted towards higher biomass production rates under warming and OA.

At higher temperatures, seawater CO<sub>2</sub> concentrations were lower than under colder conditions (Table 1; Zeebe and Wolf-Gladrow, 2001). At the same time, warming from 2 to 6 °C caused up to 60 % higher growth and 110 % higher biomass build-up rates (Fig. 1, Table 2). Furthermore, the decrease in  $\tau_{ES}$  indicates a faster transfer of photochemical energy into downstream processes such as RuBisCO activity (Table 3). Increased carbon demand in concert with lower carbon supply at higher temperatures thus increases the risk of CO<sub>2</sub> shortage in the cell, which in turn causes OA to have larger effects than at colder temperatures. Moreover, warming changes the kinetics of carbon fixation, with RuBisCO increasing its maximum turnover rates but decreasing its affinity for CO<sub>2</sub> (Young et al., 2014). At higher temperatures, cells thus have the potential for higher carboxylation rates provided sufficient CO<sub>2</sub> is available (Kranz et al., 2015). Under elevated pCO<sub>2</sub> levels, diffusive CO<sub>2</sub> supply increases and/or costs for active carbon acquisition decrease. Consequently, the positive effect of increasing temperature on the carbon fixation rate can develop its full potential under OA.

In conclusion, elevated catabolic activity under warmer conditions can explain the observed upward shift in the  $CO_2$  optimum of growth with increasing temperature (Fig. 1), as the corresponding higher carbon demand causes  $CO_2$  fixation to saturate under higher  $pCO_2$  levels. In combination with a faster and more efficient machinery for pH homeostasis at elevated temperatures (Morgan-Kiss et al., 2006), this could explain why declining growth rates were only observed at relatively higher  $pCO_2$  levels compared to those under low temperature conditions (Fig. 2).

# 4.4 Implications for the current and future Arctic pelagic ecosystem

Picoeukaryotes such as *M. pusilla* are considered to be potential winners of climate change: they are not only thriving in warmer, more stratified environments, which are predicted to further expand in the future, but also seem to benefit from OA (Li et al., 2009; Schulz et al., 2017). Our results for *M. pusilla* confirm beneficial effects of warming and OA on growth and biomass production under nutrient-replete conditions (Fig. 1, Table 2). Hence, in warmer spring conditions of the future, this species may experience growth stimulation under OA, potentially increasing its importance early in the growing season. Currently, *M. pusilla* already

dominates Arctic phytoplankton assemblages in the nutrient-limited summer and autumn situations, which were not investigated in this study. Regarding the importance of nutrient availability, laboratory experiments found beneficial OA effects on *M. pusilla* primary production to persist also under P limitation (Maat et al., 2014), while in a mesocosm community, OA-dependent increases in *M. pusilla* abundances disappeared when the system ran into P and N co-limitation (Engel et al., 2008). Thus, it remains to be seen how the combined effects of warming and OA manifest under low nutrient conditions as well as how the responses may depend on sources and types of nutrients (e.g. mixing-delivered nitrate vs. regenerated ammonium).

A species' success in the environment does not only depend on individual performance, but also on how it compares to that of competing species. When we compare our results with the responses of the Arctic diatom Thalassiosira hyalina, isolated from the same location and exposed to the same experimental conditions (Wolf et al., 2018), the diatom had higher growth rates than the picoeukaryote under most treatment conditions, as can be expected for nutrient-replete conditions (Sarthou et al., 2005). The relative increase in growth rates from ambient (2 or 3 °C and 380 µatm pCO<sub>2</sub>) to future conditions (6 °C and 1000 µatm pCO<sub>2</sub>) was, however, much higher for M. pusilla than for T. hyalina. The fact that our experiments were conducted under nutrientreplete conditions, which typically favour diatoms over picoeukaryotes, may indicate an even stronger increase in fitness (Collins et al., 2014) and could mean that M. pusilla gains another competitive advantage over phytoplankton like diatoms in the future, in addition to those resulting from changes in stratification (Li et al., 2009). Thus, our findings suggest higher picoplankton contribution to future Arctic phytoplankton assemblages under non-limiting conditions, e.g. early in the growing season when picoeukaryotes can already contribute quite substantially to the phytoplankton standing stocks (Marquardt et al., 2016; Paulsen et al., 2015). How such competition between diatoms and picoeukaryotes would manifest under nutrient-depleted conditions that strongly favour M. pusilla is currently unknown.

Even though picoeukaryotes seem to contribute more to the downward export of organic matter than previously assumed (Waite et al., 2000; Richardson and Jackson, 2007), in comparison to diatoms for example, they are less efficient vectors for carbon export to depth and have a lower energy transfer along trophic levels (Sherr et al., 2003). Consequently, Arctic food webs dominated by picoeukaryotes would look very different from those fuelled by diatom production (Sherr et al., 2003; Paulsen et al., 2015). Due to its motility and capability to grow mixotrophically, *M. pusilla* is characterized by an exceptionally high cellular C: N ratio compared to other Arctic phytoplankton (Table 2; Halsey et al., 2014; McKie-Krisberg and Sanders, 2014). An increased importance of this species would thus not only affect the food web due to its small size and concurrent grazer pref-

erences, but also in terms of food quality (van de Waal and Boersma, 2012). The expected higher growth rates and thus abundances of this species may thus strengthen the Arctic microbial food web. Together with a concurrent weakening of the classical diatom-fuelled food web, this could have severe implications for the flow of energy and nutrients through future marine Arctic ecosystems (Post, 2016).

#### 5 Conclusions

This study is the first to show synergistic effects of warming and OA on M. pusilla, one of the most abundant species of the worlds' oceans. Individually, both warming and OA cause more efficient biomass build-up under nutrient-replete conditions. Beneficial effects manifest, however, even more strongly in combination, when facilitated carbon acquisition (e.g. due to higher diffusive CO<sub>2</sub> supply) co-occurs with higher fixation rates (e.g. due to higher turnover-rates of Ru-BisCO). Our results provide an explanation for the observations of previous mesocosm studies, which indicated beneficial effects of OA and warming on M. pusilla and other picoeukaryotes. Characterising the responses of this key Arctic species to warming and OA will help to develop mechanistic phytoplankton functional types and more realistic model representation of phytoplankton assemblages as well as their responses to multiple drivers. Future studies are needed to elucidate further multifactorial environmental changes, addressing both abiotic (e.g. changes in light and nutrients) as well as biotic (e.g. heterotrophy, competition, grazers, viruses) interactions.

*Data availability.* The raw data of this study are available in the PANGAEA data repository with the identifier https://doi.org/10.1594/PANGAEA.892370 (Hoppe et al., 2018a).

The Supplement related to this article is available online at https://doi.org/10.5194/bg-15-4353-2018-supplement.

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Competing interests. The authors declare that they have no conflict of interest.

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