



High Levels of Microplastics in the Arctic Sea Ice Alga *Melosira* arctica, a Vector to Ice-Associated and Benthic Food Webs

Melanie Bergmann,* Steve Allen, Thomas Krumpen, and Deonie Allen

Cite This: Environ. Sci. Technol. 2023, 57, 6799–6807



\mathbf{a}		1
 	トく	

Metrics & More



ABSTRACT: Plastic pollution has become ubiquitous with very high quantities detected even in ecosystems as remote as Arctic sea ice and deepsea sediments. Ice algae growing underneath sea ice are released upon melting and can form fast-sinking aggregates. In this pilot study, we sampled and analyzed the ice algae*Melosira arctica*and ambient sea water from three locations in the Fram Strait to assess their microplastic content and potential as a temporary sink and pathway to the deep seafloor. Analysis by μ -Raman and fluorescence microscopy detected microplastics ($\geq 2.2 \ \mu$ m) in all samples at concentrations ranging from 1.3 to 5.7 × 10⁴ microplastics (MP) m⁻³ in ice algae and from 1.4 to 4.5 × 10³ MP m⁻³ in sea water, indicating magnitude higher concentrations in algae. On average, 94% of the total microplastic particles were identified as 10 μ m or smaller in size and comprised 16 polymer types without a clear dominance. The high concentrations of microplastics from



melting ice and ambient sea water. The algae appear to be a temporary sink and could act as a key vector to food webs near the sea surface and on the deep seafloor, to which its fast-sinking aggregates could facilitate an important mechanism of transport.

KEYWORDS: Arctic, ballasting, Fram Strait, sea ice, ice algae, Melosira arctica, microplastic, plastic, sympagic, polar regions

INTRODUCTION

Since the 1970s, plastic production has increased annually by 8% and could double in the coming 20 years.¹ Since there is still no solution for the end of life of most plastic products, plastic pollution has spread to ecosystems around the globe.² Under the influence of light, mechanical abrasion, and temperature fluctuations, plastic items break down into ever smaller fragments. Particles ≤ 5 mm are termed microplastics (MPs) and considered irretrievable.³ Despite the remoteness of the Arctic, MPs have become particularly abundant and ubiquitous in Nordic ecosystems,^{4,5} including Arctic snow, glaciers, sea water, and very high concentrations in deep-sea sediments and sea ice.^{6–10} Not surprisingly, MPs were also reported in organisms associated with sea ice such as sympagic zooplankton¹¹ and polar cod (*Boreogadus saida*).^{12,13}

The centric diatom *Melosira arctica* (Figure 1) can be considered a keystone to Arctic ecosystems feeding a variety of zooplankton species near the sea surface.¹⁴ While its extent and distribution are still poorly known,¹⁵ it is an important primary producer, which accounted for 45% of the Arctic primary production in 2012,¹⁶ the year of the lowest hitherto recorded Arctic sea ice extent. This unicellular microalga forms colonial filamentous strands, nets, ropes, mats, and curtains of up to 3 m length underneath annual and multi-year sea ice and is characterized by a patchy distribution.^{15,17,18} During sea ice melting, *M. arctica* detaches from the sea ice, and its aggregates

float freely^{15,19} or sink rapidly to the ocean floor (Figure 1A–C), feeding epibenthic organisms such as brittlestars and sea cucumbers.^{16,17,20} *M. arctica* is widely distributed with reports from the Canada Basin, Chukchi Sea, Nunavut, northern Baffin Bay, Arctic Ocean, Central Arctic, Laptev Sea, northeast Greenland, and the Barents Sea.^{15–17,20}

Given its sticky filamentous morphology and its close association with Arctic sea ice, which carries very high levels of MPs, we hypothesize that *M. arctica* entraps MPs released from the overlying melting sea ice and surrounding sea water. To test this hypothesis, we analyzed *M. arctica* samples that were taken opportunistically during helicopter-based ice stations of a research campaign in 2021 at the HAUSGARTEN observatory in the Fram Strait. Satellite-based backtracking of the sampled ice floes was undertaken to pinpoint the origin and trajectory of the ice floes and potentially also of the entrained MPs. It is important to fill this knowledge gap since*M. arctica* could function as a vector of MPs into sub-ice and benthic food webs. Given the importance of*M. arctica* in Arctic benthic food webs, it

Received:October 28, 2022Revised:March 16, 2023Accepted:March 16, 2023Published:April 21, 2023







Figure 1. (A) Photographs of *M. arctica* in sea water and broken ice surrounding the ice floes at sites Mel_2 and (B) Mel_3. (C) Microscopic photograph of a subsample of Mel_1 taken for the species identification of *M. arctica* (credit with permission: A. Kraberg, AWI, magnification: ×400). (D) Ice algae (black arrows) deposited on the seafloor at ~2500 m depth in the Fram Strait, depicted by a time-lapse camera on 24/07/20 (credit with permission: F. Wenzhöfer, AWI). (E) Map of the study area and sampling sites in the Fram Strait; red arrows indicate water masses of the Atlantic inflow, and blue arrows depict cooler waters of polar origin (current arrows are inspired with permission by Macdonald et al.²¹).

Table 1. Details of M. arctica and Se	a Water Sampling
---------------------------------------	------------------

station	time (UTC)	date	position latitude	position longitude	ice salinity (EC mS cm ⁻¹)	ice floe thickness (m)
Mel_1	13:00 h	13/06/21	78.90194 N	3.151111 W	3.78-35.4	2.13
Mel_2	15:00 h	14/06/21	78.59433 N	3.548667 W	3.38-23.6	2.02
Mel_3	13:00 h	15/06/21	79.02889 N	5.713889 W	14.26-23.5	1.52

is important that we understand the likelihood of *M. arctica* acting as a scavenger and transport vector for MP and plastic chemicals.

MATERIALS AND METHODS

Sampling Procedure. Three *M. arctica* samples were collected in the Arctic circle in 2021 during expedition PS121 of the research icebreaker RV Polarstern to the HAUSGARTEN observatory, which saw substantial quantities of ice floes. Samples were collected from sea water adjacent to the floating sea ice (Table 1, Figure 1E).

Ice floes were visited by a helicopter (D-HAOE, D-HAPS, HeliService, Emden, Germany) during a sampling campaign for MPs in snow, melt ponds, sea water, and sea ice and had to be safe for landing. Here, we present only the results from M. arctica and corresponding sea water samples as all other analyses are still ongoing. Bending over the ice edge (secured by a safety rope), M. arctica aggregates that were free-floating among the sea ice (Figure 1A,B) were scooped from the sea water using a nylon net of ~1 mm porosity and placed into triple MilliQ pre-rinsed 2 L stainless-steel containers (ECOtanka). Sea water samples were also grabbed from the sea surface with triple MilliQ pre-rinsed 2 L stainless-steel containers attached to a hemp rope (4 L in total). Field blank samples for each location were created using 2 L stainless-steel containers left open for the duration of sampling with 100 mL of MilliQ water in the base to retain deposited MP particles. Scientists stood downwind of the sampling location when taking the sample to minimize any clothing contamination since red survival suits and leather gloves had to be worn for work safety. The position of the ice floe was recorded at the end of the visit by a hand-held GPS device (Table 1).

Sample Preparation. Back on the research vessel, ice algal samples were transferred into triple MilliQ-rinsed glass vials and refrigerated (at 4 °C) until analysis. A subsample was taken for species identification under the microscope (Figure 1C). The sea water samples were filtered onto Whatman QMA quartz filters (2.2 μ m porosity, liquid) and stored at room temperature in sealed metal containers (20 °C) in the laboratory. Organic matter was digested from sea water filtered samples using hydrogen peroxide⁷ (40 mL of H_2O_2 was placed over the filtered samples held in borosilicate glass filtration equipment and incubated at room temperature for 48 h and then flushed with 200 mL of MilliQ). Then, the air-dried filter was stored in sealed metal containers until analysis. In a dedicated MP laboratory, M. *arctica* samples were subsampled for MP and nanoplastic (NP) analysis in a laminar flow fume hood that had been pre-cleaned to minimize contamination. Using sterilized stainless-steel tweezers and sterilized stainless-steel micro spatulas, subsamples of 3-7 g of wet material (see Supplementary Data) were placed into sterilized borosilicate glass test tubes (four subsamples per site). 20 mL of 30% w/w hydrogen peroxide (H_2O_2) , pre-filtered through a 0.2 μ m aluminum oxide filter) was added to each sample with sterilized aluminum foil lids, and then, the samples were placed in a dry heat block for 48 h at 50 °C. An additional 10 mL of H_2O_2 was then added, and the samples were returned to the dry heat block for another 48 h at 50 °C. The samples were then filtered onto 47 mm diameter aluminum oxide filters (0.2 μ m porosity) and flushed with 250 mL of MilliQ for Raman analysis. Filters did not become blind during any of the collection or filtration processes. The filters were stored in sterilized 50 mm diameter aluminum containers to minimize contamination.

Raman Analysis. All samples were then analyzed by μ -Raman microscopy following published protocols developed by Zhao et al.²² and used consistently by Allen et al.^{23–25} (Horiba Xplora Plus) using a 785 nm laser (spatial resolution of 1 μ m), 1200 gr mm⁻¹ grating, 50 μ m slit, and 25% power (filter), with spectra collected over 200–2000 cm⁻¹ Raman shift using five acquisitions of 10 s. 25% of each filter was assessed following the filter area representation of Huppertsberg and Knepper.²⁶ Raman spectra were analyzed using Spectragryph,²⁷ SloPP, SloPP-E,²⁸ and siMPle²⁹ Raman reference libraries supplemented with in-house plastic spectra references. Spectra with a \geq 80% hit rate were counted as plastic particles and included in the total sample MP count.^{23,24,30–33} The limit of quantification (LOQ, s. below) for this analysis was set to 1.2 μ m for *M. arctica* samples and 2.2 μ m for sea water samples. This was due to the different porosity of filtration possible for the two different matrices. The μ -Raman results were used to quantify the number of MP particles and the quantity of different plastic polymer types in each sample according to Armbruster and Pry^{34} and Dawson et al.,³⁵ where $LOQ \ge LOD$ (Limit of Detection) and LOD = $(\text{mean}_{\text{blank}} + 1.645 \text{ (SD}_{\text{blank}})) + 1.645$ (SD_{lowest sample result}).

Fluorescence Microscopy. Samples were secondarily analyzed using Nile Red fluorescence microscopy³⁶⁻³⁸ to provide a particle count check and the particle shape and size representative of each sample. Particles were identified as fibers $(1:3 \text{ width-to-length ratio})^{39,40}$ or fragments due to the uncertainty in defining films. Samples were stained with Nile Red powder (Sigma Aldrich) dissolved in methanol (99% pure, Sigma Aldrich) to result in 0.1 μ g mL⁻¹ solution and pre-filtered through sterilized glass fiber filters (Whatman GF/C, 1.2 μ m porosity).³⁷ The Nile Red solution was stored in a sterilized, sealed glass container covered in aluminum foil in the dark at 4 °C. Stained samples were oven-dried (30 °C) overnight and then placed in the fridge $(4 \,^{\circ}C)$ in sealed aluminum containers to rest prior to fluorescence microscopy imaging. A Nikon LV100ND fluorescence microscope with an FITC (Fluorescein) filter was used to image 25% of each filter following the filter area representation.²⁶ Fluorescent particle dimensions (including circularity and the ferret diameter) and counts were collected using FIJI software^{37,41} following published standards.^{41,42} Nile Red fluorescence MP counts were used to confirm the Raman MP quantification and to obtain particle size and shape information.

Blanks, Positive Controls, and Contamination Mitigation. Sample preparation was undertaken in a dedicated MP laboratory with controlled access, a dedicated HEPA filter, and a laminar flow fume hood. Prior to sample preparation, the laboratory and laboratory fume hood were pre-cleaned to minimize laboratory contamination. Pre-cleaning was completed using a plastic-free cleaning fluid followed by MilliQ water and clean 100% cotton lint-free clothes. Only cotton clothing and cotton lab coats were worn in the laboratory, nitrile gloves were used for all chemical handling, and all materials were sterilized (superheated to 450 °C) or triple-washed with MilliQ (as appropriate) between each use.

Field blank samples underwent the sample digestion process and full analytical process, resulting in all field blanks being full process blanks. All samples were blank-corrected using these field blank μ -Raman results. Blanks had on average 5 (\pm 2) particles per sampled filter area (<20% of the analyzed field MP particle results per total filter count and therefore supporting field result validation). Positive controls (laboratory blanks



Figure 2. MP quantities, particle size distribution, and polymer types detected in the samples of the ice alga *M. arctica*. (A) Boxplot of MP particle counts (10th–90th percentile), means, and standard deviations in *M. arctica* and surrounding sea water samples (MP m⁻³). (B) *M. arctica* MP counts (10th–90th percentile), means, and standard deviations relative to the sample wet and dry weights (g). The asterisks indicate datasets that present a significant difference (* = p < 0.05, ** = p < 0.01).

spiked with known plastic particles) were also created to check the full process recovery efficiency. A spiked solution was made from polyethylene particles ranging from 2 to 110 μ m (particle size distribution described using a Mastersizer 2000), and spiked solution samples underwent the full preparation and analysis process as field samples and field blanks. The μ -Raman analytical efficiency (the effective polymer counts relative to this analytical method quantified by positive controls) was calculated to be 89% (\pm 12%) and the Nile Red fluorescence microscopy analysis efficiency to be 86% (\pm 12%) (see the Supplementary Information). It is acknowledged that spiked recovery samples should ideally include a range of polymer types, and future studies will incorporate this advancement.

Backtracking of Ice Floes Surrounding Ice Algae. The possible origin and pathways of the ice algal samples were determined using two different satellite-based tracking approaches: in the Fram Strait, floe locations were tracked visually using optical satellite data. Visual identification and tracking were carried out using NASA's interactive interface to display full-resolution satellite images on a daily basis (https:// worldview.earthdata.nasa.gov/). After the floes reached the closed ice cover at ~82°N, tracking was continued using an automated backtracking algorithm based on passive microwave data. Note that the automated tracking is less accurate in the Fram Strait,⁴³ which is why we have limited its use to the central Arctic. The applied automated backtracking algorithm, products, and their accuracy are described in Krumpen et al.⁴⁴ Tracking was discontinued if the ice concentration at a specific location along the backward trajectory dropped below 20% and we assumed the ice to be formed.

Data Analysis. Nile Red fluorescence microscopy analysis was undertaken using FIJI(ImageJ)^{37,41} exported into standard Excel software for blank correction and result plotting. Similarly, Raman spectral analysis was undertaken using Spectragryph²⁷ software with the results exported into standard Excel software for blank correction and result plotting. Blank correction was undertaken by the particle size (for particle size distribution analysis) and the polymer type (for polymer sample composition analysis). Blank correction of all samples was undertaken using the full process field blank sample results to ensure that all field and laboratory actions were accounted for in

contamination assessment and the most conservative approach to blank correction was used. Simple correlation and statistical tests were completed using Excel. The underlying data and test results are available in the Supplementary Information.

RESULTS AND DISCUSSION

MP particles were found in all subsamples across a variety of polymer types and particle sizes. A total of 400 MP particles were identified from the 12 subsamples (Figure 2A and Supplementary Information). M. arctica samples contained 5-66 MP particles per mg wet weight (Figure 2B), on average 28 (± 20 standard deviation, σ) MP particles per sample, and an estimated equivalent mean of $3.1 \pm 1.9 \times 10^4$ MP m⁻³ (see Table S1 for sample MP counts, weights, and volumes and resulting MP values per mg and m³). While sea ice in this region contained higher MP levels $(1.2 \times 10^7 - 1.1 \times 10^6 \text{ MP m}^{-3})$,⁷ sea ice concentrates suspended organic carbon by two orders of magnitude relative to ambient sea water during ice formation through a process termed "scavenging", which partly explains the exceptionally high MP concentrations (suspended organic carbon in Arctic sea ice may also be two orders of magnitude greater than that in sea water).⁴⁵ If two orders of magnitude are deducted from the MP concentrations for a better comparison, the MP concentrations of sea ice and ice algae become strikingly similar at the magnitude scale despite differences in the methodology used. MP concentrations in water samples ranged from 4.5 to 1.4×10^3 MP m⁻³ with an overall mean of $2.8 \pm 1.3 \times$ 10^3 MP m^{-3} .

There were statistical differences in MP concentrations for the MP count relative to wet weight (g) and volume (MP m⁻³) (Kruskal–Wallis one-way analysis of variance, df = 2, H = 8.436, and p = 0.0154) with significantly highest levels found in the easternmost sample Mel_1 ($5.7 \pm 1.7 \times 10^4$ MP m⁻³), followed by the westernmost Mel_3 ($2.2 \times 10^4 \pm 8.1 \times 10^3$ MP m⁻³) and nearby Mel_2 ($1.3 \times 10^4 \pm 8.7 \times 10^3$ MP m⁻³, Figure 2A). Interestingly, the concentrations of MP in the sea water samples followed a similar trend to *M. arctica* samples from the respective sites with the highest levels found in Mel_1 ($4.5 \pm 1.5 \times 10^3$ MP m⁻³) followed by Mel_3 ($2.5 \times 10^3 \pm 2.8 \times 10^2$ MP m⁻³) and Mel_2 ($1.4 \pm 1.0 \times 10^3$ MP m⁻³, Figure 2A). A pairwise *t*-test for independent means showed that *M. arctica* samples contained

significantly higher MP levels compared with their respective sea water samples ($t_{(9)} = -3.75$, p = 0.006 for total sea water MP; $t_{(9)} = 4.046$, p = 0.003 for sea water MP < 25 μ m). This could be interpreted as an indication of *M. arctica* trapping MPs from the surrounding sea and melt water during ice drift (Figure 3). While



Figure 3. Drift trajectory of ice floes sampled pointing to formation in the Laptev Sea. Solid lines result from an automated sea ice tracking algorithm, while dashed lines are derived from a visual analysis of optical satellite data in the Fram Strait area.

the water samples represent only snapshots in time, *M. arctica* could have trapped MPs from the surrounding water over time and thus accumulated MPs, leading to the higher concentrations measured. Interestingly, the sample Mel_2 with the lowest concentration floated near the remains of an ice berg (glacial origin, Figure 1A), whereas the other samples came from sea ice floes.

Since the ice berg (Mel_2) could not be located on highresolution satellite data, suggesting that the source was unknown or that it drifted with the sea ice floes (Mel_1), it was combined with the latter's drift trajectory (Figure 3). The paths of all ice floes sampled looked remarkably similar and suggest that the ice was 560 days old and formed off the Laptev Sea in mid-October (Figure 3). However, this origin does not necessarily reflect the origin of the ice algae. M. arctica can become detached due to strong currents and re-attach itself using extracellular polymeric substances.¹⁵ Therefore, this association could be quite dynamic. Still, the drift path taken is quite representative of sea ice in this area and could indicate where along the way MPs became entrained in the algal matrix. The MP concentrations in sea water $(0-6.3 \times 10^3 \text{ MP m}^{-3})$ were higher than previous measurements from the Fram Strait $(0-1287 \text{ MP m}^{-3})$,⁹ which is likely due to differences in the methodology used (32 μ m pore size of pump filters compared to the 2.2 μ m pore size limit in this study, μ -FTIR analysis, and larger water volume sampled in⁹). Constraining the sea water sample results to the particle size limit,⁹ our mean sea water samples for particles >25 μ m (527 ± 294 MP m^{-3}) are comparable to those previously published $(421 \text{ MP m}^{-3}).$

MP particles in both algal and sea water samples were generally smaller than 40 μ m, with on average 94% (88–97%) of the total MP particles identified as 10 μ m or smaller in size (Figure 4). Sample Mel_3, the most western site closest to the cold south-flowing Transpolar Drift current,²¹ presented a greater quantity of smaller MP particles (69% <5 μ m), while the more eastern samples contained a slightly larger proportion of >5 μ m particles (59–60%).

The *M. arctica* cell size ranges from 20 to 30 μ m.^{16,20} Thus, these small MP particles could have been entangled in the filamentous threads, e.g., "stuck" to the outside of the algal aggregates, a process that could be helped by the mucous extracellular polymeric polysaccharides produced by the algae.^{15,46} Suspended MPs were also observed to adhere to the



Figure 4. Characteristics of identified MPs. (A) Photographs of MPs. (B) Overarching and *M. arctica* sample MP particle size distribution. The asterisk indicates datasets that present a significant difference (* = p < 0.05). (C) Individual and overall polymer composition. Polymer composition within the samples varied but included polyethylene (PE), polyester (PES), polyethylene terephthalate (PET), polypropylene (PP), polystyrene (PS), polyvinyl chloride (PVC), polyurethane (PU), acrylonitrile butadiene styrene (ABS), polycarbonate (PC), polyethylene vinyl acetate (PVA), nylon, polyamide (PA), acrylic, cellulose acetate (CE), polymethyl methacrylate (PMMA), polysulfone (PLS), rubber, polylactic acid (PLA), styrene butadiene (SBT), styrene acrylonitrile copolymer (SAN), and others (<5%).

attenuating light,⁴⁸ a process that has been reported for several algal species.⁴⁹ Small MPs could also have occurred within the *M. arctica* cells. While differentiation was beyond the scope of this pilot study, it has been suggested that small MP particles could damage and enter algal cells, leading to oxidative stress responses including the damage of chloroplasts and thus inhibition of photosynthesis depending on the experimental doses used.⁴⁹ Both mechanisms could have serious implications in terms of carbon sequestration and the ocean's capacity to mitigate climate change, especially as MP concentrations are set to increase at the very least due to fragmentation of plastic debris that is already in the ocean.³

The occurrence and particle size distribution of these small MPs suggest that *M. arctica* could be entraining these smaller MP particles from the surrounding sea water and the melting sea ice. Sea ice is reported to contain MP particles up to $50 \,\mu\text{m}$ with a notable percentage of particles sized $11-25 \,\mu\text{m}$,⁷ a larger range of particle sizes and a greater quantity of slightly larger particles than those found in the *M. arctica* samples. Similarly, snow and Arctic sea water are reported to present a larger MP range and particle size (up to $450 \,\mu\text{m}$ in snow and $\leq 200 \,\mu\text{m}$ in sea water).⁹

Despite the small sample size, the MP particles comprised 16 different polymer types without a clear dominance of any single polymer type (Figure 4). Similarly, Arctic sea ice, which is closely linked with ice algae, contained 17 different polymer types.⁴ PET accounted for the highest proportion of particles overall (13%). Polymers commonly used in paints and building/ construction materials (acrylics and polymethyl methacrylate, PMMA) formed the second and third highest proportions of the particles (12 and 10%, respectively), concurring with elevated paint-related particles found in Arctic snow and water samples.^{6,9} Polyester (PES), polyethylene (PE), and polyamide (PA) were also notable in all samples (9, 9, and 7%, respectively, overall). Polyester and polyamide have been reported in Arctic sea water in high quantities (up to 39%).^{8,9,50} Similarly, polyvinyl chloride (PVC) was also reported in deep-sea sediments (1-13%),⁵¹ sea water (<30%),⁵⁰ and sea ice (2-5%)^{7,50} from the Arctic. There is also an atmospheric component to the Arctic MP cycle,^{5,10,52} and atmospheric MPs are known to comprise a wide range of polymer types that can be rapidly transported over vast distances helped by resuspension of MPs from the sea surface.^{25,53} Both the polymer composition and the particle size distribution found in the M. arctica samples indicate that atmospheric deposition could play a role along with sea water transported into and out from the Arctic and from sea ice and snow melt.

Possible Influence of Climate Change and *M. arctica* **MP in the Arctic Food Chain.** *M. arctica* is a primary producer that colonizes the underside of Arctic sea ice, growing during the summer periods when photosynthesis is possible due to available sunlight and nutrients.⁵⁴ As the sea ice melts, the detached *M. arctica* is released into the water, where it produces high amounts of extracellular polymeric substances and forms aggregates.^{46,55} Such aggregates can stay afloat if irradiance promotes photosynthesis, and the resulting oxygen bubbles are captured in the viscous aggregates.⁵⁵ During the next freeze-up, this could seed the new-forming ice and form the basis of new *M. arctica* colonies,¹⁵ a process that could also inoculate new-forming ice with MPs.

However, an important fraction of the *M. arctica* aggregates sinks to the Arctic seafloor, which accounted for 85% of the vertical carbon export in the Arctic in 2012.¹⁶ Katlein et al.¹⁵ suggested that aggregates with a diameter of 3 cm can reach the deep seafloor within a timeframe as short as 1 day. Bergmann et al.⁵ suggested that MPs bound to ice algal aggregates could contribute to the vertical export of MPs from the sea surface to the seafloor since concentrations of 6.3×10^6 MP m⁻³ were measured in the deep-sea sediments beneath the marginal ice zone and a positive correlation with the chlorophyll *a* content was found.⁹

Following the current trend in climate change, resulting in hotter temperatures and more extreme weather events, sea ice will become thinner and less abundant^{54,56} and break up increasingly because of the increased storm frequency, potentially supporting greater ice algal blooms due to higher light availability. However, this is likely offset by nutrient limitation and the shorter frozen periods, with the ice habitat melting earlier and freezing up later than in past decades, resulting in shorter algal growth periods⁵⁷ as well as lower biomass and deposition. If M. arctica is a key transport vector of MPs from the sea surface to the seafloor, then any shift in M. arctica growth and circulation will also modify the transport and distribution of MPs. However, other phytoplankton species, such as Phaeocystis pouchetii, which can form extensive floating gelatinous colonies in Arctic waters that also produce sticky exopolymer particles⁵⁸ and sink to the seafloor,⁵⁹ could play this role.

No matter what the consequences, it could be speculated that MPs from melting sea ice could be trapped by Arctic ice algae as a result of its filamentous and sticky morphology. Since ice algae are hotspots of biological activity and an important food source for grazing organisms,^{14,19,60} they could be a vector into underice food webs. Indeed, MPs were recently detected in grazing zooplankton from the Fram Strait.¹¹ Similarly, it could pave the uptake of MPs by benthic organisms that have been observed to feed on ice algae deposited on the seafloor.^{16,61} While the effect of this uptake on Arctic species and ecosystems is not yet known,⁵ Arctic biota are already under serious pressure from global heating, which progresses four times faster in the Arctic compared with the globe.⁶² Plastic pollution likely exacerbates this pressure, so it needs to be tackled efficiently.⁶³

ASSOCIATED CONTENT

③ Supporting Information

The Supporting Information is available free of charge at https://pubs.acs.org/doi/10.1021/acs.est.2c08010.

MP counts in *M. arctica* and sea water samples from different locations; polymer types as percentage of the total MP count detected in *M. arctica* and sea water; particle size distribution of MPs in *M. arctica* samples; results of Kruskal–Wallis and *t*-tests and correlation and linear regression analyses; and results of positive control tests (PDF)

AUTHOR INFORMATION

Corresponding Author

Melanie Bergmann – HGF-MPG Group for Deep Sea Ecology and Technology, Alfred-Wegener-Institut Helmholtz-Zentrum für Polar- und Meeresforschung, 27570 Bremerhaven, Germany; orcid.org/0000-0001-5212-9808; Email: Melanie.Bergmann@awi.de

Environmental Science & Technology

Authors

- Steve Allen Ocean Frontiers Institute, Dalhousie University, B3H 4R2 Nova Scotia, Canada; Ocid.org/0000-0002-2333-6514
- **Thomas Krumpen** Sea Ice Physics, Alfred-Wegener-Institut Helmholtz-Zentrum für Polar- und Meeresforschung, 27570 Bremerhaven, Germany
- Deonie Allen School of Geography, Earth and Environmental Science, University of Birmingham, B15 2TT Birmingham, U. K.; School of Physical and Chemical Sciences, University of Canterbury, 8041 Christchurch, New Zealand; orcid.org/ 0000-0002-4038-9394

Complete contact information is available at: https://pubs.acs.org/10.1021/acs.est.2c08010

Author Contributions

M.B.: writing, editing, study design, and sampling; D.A.: writing, editing, study design, sampling, laboratory analysis, and statistical analysis; S.A.: writing, editing, laboratory analysis, and statistical analysis; and T.K.: ice floe modeling, writing, and editing.

Notes

The authors declare no competing financial interest.

ACKNOWLEDGMENTS

We gratefully acknowledge the principal scientist, officers, and crew of RV Polarstern cruise PS126 as well as the Deutsche Wetter Dienst and HeliService crews. A. Kraberg helped with the identification of the ice algae. M. Hoppmann, D. Scholz, S. Escalle, and K. Mettfies assisted during ice stations. This work was supported by the Leverhulme Trust through grant ECF-2019-306 (D.A.) and the Ocean Frontiers Institute (S.A.). This project has received funding from the European Union's Horizon 2020 research and innovation program under the Marie Skłodowska-Curie grant agreement no. 101023635 (D.A.). This work also contributes to the tasks of the Pollution Observatory of the Helmholtz Association-funded program FRAM (Frontiers in Arctic Marine Research).

REFERENCES

(1) Geyer, R. A Brief History of Plastics. In *Mare Plasticum - The Plastic Sea: Combatting Plastic Pollution Through Science and Art,* Streit-Bianchi, M.; Cimadevila, M.; Trettnak, W., Eds.; Springer International Publishing: Cham, 2020; pp. 31–47.

(2) Tekman, M. B.; Walther, B. A.; Peter, C.; Gutow, L.; Bergmann, M. Impacts of Plastic Pollution in the Oceans on Marine Species, Biodiversity and Ecosystems; WWW Germany: Berlin, 2022.

(3) MacLeod, M.; Arp, H. P. H.; Tekman, M. B.; Jahnke, A. The global threat from plastic pollution. *Science* **2021**, *373*, 61–65.

(4) Lima, A. R. A.; Ferreira, G. V. B.; Barrows, A. P. W.; Christiansen, K. S.; Treinish, G.; Toshack, M. C. Global patterns for the spatial distribution of floating microfibers: Arctic Ocean as a potential accumulation zone. *J. Hazard. Mater.* **2021**, *403*, No. 123796.

(5) Bergmann, M.; Collard, F.; Fabres, J.; Gabrielsen, G. W.; Provencher, J. F.; Rochman, C.; Van Sebille, E.; Tekman, M. B. Plastic pollution in the Arctic. *Nat. Rev. Earth Environ.* **2022**, *3*, 323–337.

(6) Bergmann, M.; Mützel, S.; Primpke, S.; Tekman, M. B.; Trachsel, J.; Gerdts, G. White and wonderful? Microplastics prevail in snow from the Alps to the Arctic. *Sci. Adv.* **2019**, *5*, No. eaax1157.

(7) Peeken, I.; Primpke, S.; Beyer, B.; Gütermann, J.; Katlein, C.; Krumpen, T.; Bergmann, M.; Hehemann, L.; Gerdts, G. Arctic sea ice is an important temporal sink and means of transport for microplastic. *Nat. Commun.* **2018**, *9*, 1505.

(8) Lusher, A. L.; Tirelli, V.; O'Connor, I.; Officer, R. Microplastics in Arctic polar waters: the first reported values of particles in surface and sub-surface samples. *Sci. Rep.* **2015**, *5*, 14947.

(9) Tekman, M. B.; Wekerle, C.; Lorenz, C.; Primpke, S.; Hasemann, C.; Gerdts, G.; Bergmann, M. Tying up loose ends of microplastic pollution in the Arctic: Distribution from the sea surface, through the water column to deep-sea sediments at the HAUSGARTEN observatory. *Environ. Sci. Technol.* **2020**, *54*, 4079–4090.

(10) Materić, D.; Kjær, H. A.; Vallelonga, P.; Tison, J.-L.; Röckmann, T.; Holzinger, R. Nanoplastics measurements in Northern and Southern polar ice. *Environ. Res.* **2022**, *208*, No. 112741.

(11) Botterell, Z. L. R.; Bergmann, M.; Hildebrandt, N.; Krumpen, T.; Steinke, M.; Thompson, R. C.; Lindeque, P. K. Microplastic ingestion in zooplankton from the Fram Strait in the Arctic. *Sci. Total Environ.* **2022**, *831*, No. 154886.

(12) Kühn, S.; Schaafsma, F. L.; van Werven, B.; Flores, H.; Bergmann, M.; Egelkraut-Holtus, M.; Tekman, M. B.; van Franeker, J. A. Plastic ingestion by juvenile polar cod (*Boreogadus saida*) in the Arctic Ocean. *Polar Biol.* **2018**, *41*, 1269–1278.

(13) Morgana, S.; Ghigliotti, L.; Estévez-Calvar, N.; Stifanese, R.; Wieckzorek, A.; Doyle, T.; Christiansen, J. S.; Faimali, M.; Garaventa, F. Microplastics in the Arctic: A case study with sub-surface water and fish samples off Northeast Greenland. *Environ. Pollut.* **2018**, *242*, 1078–1086.

(14) Leu, E.; Søreide, J. E.; Hessen, D. O.; Falk-Petersen, S.; Berge, J. Consequences of changing sea-ice cover for primary and secondary producers in the European Arctic shelf seas: Timing, quantity, and quality. *Prog. Oceanogr.* **2011**, *90*, 18–32.

(15) Katlein, C.; Fernández-Méndez, M.; Wenzhöfer, F.; Nicolaus, M. Distribution of algal aggregates under summer sea ice in the Central Arctic. *Polar Biol.* **2015**, *38*, 719–731.

(16) Boetius, A.; Albrecht, S.; Bakker, K.; Bienhold, C.; Felden, J.; Fernández-Méndez, M.; Hendricks, S.; Katlein, C.; Lalande, C.; Krumpen, T.; Nicolaus, M.; Peeken, I.; Rabe, B.; Rogacheva, A.; Rybakova, E.; Somavilla, R.; Wenzhöfer, F. Export of algal biomass from the melting arctic sea ice. *Science* **2013**, *339*, 1430–1432.

(17) Ambrose, W. G.; Quillfeldt, C. V.; Clough, L. M.; Tilney, P. V. R.; Tucker, T. The sub-ice algal community in the Chukchi sea: large- and small-scale patterns of abundance based on images from a remotely operated vehicle. *Polar Biol.* **2005**, *28*, 784–795.

(18) Hegseth, E. N.; von Quillfeldt, C. The Sub-Ice Algal Communities of the Barents Sea Pack Ice: Temporal and Spatial Distribution of Biomass and Species. *J. Mar. Sci. Eng.* **2022**, *10*, 164.

(19) Assmy, P.; Ehn, J. K.; Fernández-Méndez, M.; Hop, H.; Katlein, C.; Sundfjord, A.; Bluhm, K.; Daase, M.; Engel, A.; Fransson, A.; Granskog, M. A.; Hudson, S. R.; Kristiansen, S.; Nicolaus, M.; Peeken, I.; Renner, A. H. H.; Spreen, G.; Tatarek, A.; Wiktor, J. Floating Ice-Algal Aggregates below Melting Arctic Sea Ice. *PLoS One* **2013**, *8*, No. e76599.

(20) Poulin, M.; Underwood, G. J. C.; Michel, C. Sub-ice colonial *Melosira arctica* in Arctic first-year ice. *Diatom Res.* **2014**, *29*, 213–221.

(21) Macdonald, R. W.; Harner, T.; Fyfe, J. Recent climate change in the Arctic and its impact on contaminant pathways and interpretation of temporal trend data. *Sci. Total Environ.* **2005**, *342*, 5–86.

(22) Zhao, S.; Danley, M.; Ward, J. E.; Li, D.; Mincer, T. J. An approach for extraction, characterization and quantitation of microplastic in natural marine snow using Raman microscopy. *Anal. Methods* **2017**, *9*, 1470–1478.

(23) Allen, S.; Allen, D.; Baladima, F.; Phoenix, V. R.; Thomas, J. L.; Le Roux, G.; Sonke, J. E. Evidence of free tropospheric and long-range transport of microplastic at Pic du Midi Observatory. *Nat. Commun.* **2021**, *12*, 7242.

(24) Allen, S.; Allen, D.; Phoenix, V. R.; Le Roux, G.; Durántez Jiménez, P.; Simonneau, A.; Binet, S.; Galop, D. Atmospheric transport and deposition of microplastics in a remote mountain catchment. *Nat. Geosci.* **2019**, *12*, 339–344.

(25) Allen, S.; Allen, D.; Moss, K.; Le Roux, G.; Phoenix, V. R.; Sonke, J. E. Examination of the ocean as a source for atmospheric microplastics. *PLoS One* **2020**, *15*, No. e0232746.

(26) Huppertsberg, S.; Knepper, T. P. Instrumental analysis of microplastics—benefits and challenges. *Anal. Bioanal. Chem.* **2018**, 410, 6343–6352.

(27) Menges, F. Spectragryph - optical spectroscopy software Version 1.2.14. http://www.effemm2.de/spectragryph/ (accessed in 2021).

(28) Munno, K.; De Frond, H.; O'Donnell, B.; Rochman, C. M. Increasing the Accessibility for Characterizing Microplastics: Introducing New Application-Based and Spectral Libraries of Plastic Particles (SLOPP and SLOPP-E). *Anal. Chem.* **2020**, *92*, 2443–2451.

(29) Primpke, S.; Cross, R. K.; Mintenig, S. M.; Simon, M.; Vianello, A.; Gerdts, G.; Vollertsen, J. Toward the Systematic Identification of Microplastics in the Environment: Evaluation of a New Independent Software Tool (siMPle) for Spectroscopic Analysis. *Appl. Spectrosc.* **2020**, *74*, 1127–1138.

(30) Lenz, R.; Enders, K.; Stedmon, C. A.; Mackenzie, D. M. A.; Nielsen, T. G. A critical assessment of visual identification of marine microplastic using Raman spectroscopy for analysis improvement. *Mar. Pollut. Bull.* **2015**, *100*, 82–91.

(31) Khashaba, P. Y.; Ali, H. R. H.; El-Wekil, M. M. A rapid Fourier transform infrared spectroscopic method for analysis of certain proton pump inhibitors in binary and ternary mixtures. *Spectrochim. Acta A Mol. Biomol.* **2018**, *190*, 10–14.

(32) Ševčík, R.; Mácová, P. Localized quantification of anhydrous calcium carbonate polymorphs using micro-Raman spectroscopy. *Vib. Spectrosc.* **2018**, *95*, 1–6.

(33) Lagaron, J. M.; Dixon, N. M.; Reed, W.; Pastor, J. M.; Kip, B. J. Morphological characterisation of the crystalline structure of colddrawn HDPE used as a model material for the environmental stress cracking (ESC) phenomenon. *Polymer* **1999**, *40*, 2569–2586.

(34) Armbruster, D. A.; Pry, T. Limit of blank, limit of detection and limit of quantitation. *Clin. Biochem. Rev.* **2008**, *29*, S49–S52.

(35) Dawson, A. L.; Santana, M. F. M.; Nelis, J. L. D.; Motti, C. A. Taking control of microplastics data: A comparison of control and blank data correction methods. *J. Hazard. Mater.* **2023**, 443, No. 130218.

(36) Shim, W. J.; Song, Y. K.; Hong, S. H.; Jang, M. Identification and quantification of microplastics using Nile Red staining. *Mar. Pollut. Bull.* **2016**, *113*, 469–476.

(37) Erni-Cassola, G.; Gibson, M. I.; Thompson, R. C.; Christie-Oleza, J. A. Lost, but Found with Nile Red: A Novel Method for Detecting and Quantifying Small Microplastics (1 mm to 20 μ m) in Environmental Samples. *Environ. Sci. Technol.* **2017**, *51*, 13641–13648.

(38) Nel, H. A.; Chetwynd, A. J.; Kelleher, L.; Lynch, I.; Mansfield, I.; Margenat, H.; Onoja, S.; Goldberg Oppenheimer, P.; Sambrook Smith, G. H.; Krause, S. Detection limits are central to improve reporting standards when using Nile red for microplastic quantification. *Chemosphere* **2021**, *263*, No. 127953.

(39) Cole, M. A novel method for preparing microplastic fibers. *Sci. Rep.* **2016**, *6*, 34519.

(40) Kooi, M.; Koelmans, A. A. Simplifying Microplastic via Continuous Probability Distributions for Size, Shape, and Density. *Environ. Sci. Technol. Lett.* **2019**, *6*, 551.

(41) Schindelin, J.; Arganda-Carreras, I.; Frise, E.; Kaynig, V.; Longair, M.; Pietzsch, T.; Preibisch, S.; Rueden, C.; Saalfeld, S.; Schmid, B.; Tinevez, J.-Y.; White, D. J.; Hartenstein, V.; Eliceiri, K.; Tomancak, P.; Cardona, A. Fiji: an open-source platform for biological-image analysis. *Nat. Methods* **2012**, *9*, 676–682.

(42) Cowger, W.; Gray, A.; Christiansen, S. H.; DeFrond, H.; Deshpande, A. D.; Hemabessiere, L.; Lee, E.; Mill, L.; Munno, K.; Ossmann, B. E.; Pittroff, M.; Rochman, C.; Sarau, G.; Tarby, S.; Primpke, S. Critical Review of Processing and Classification Techniques for Images and Spectra in Microplastic Research. *Appl. Spectrosc.* **2020**, *74*, 989–1010.

(43) Krumpen, T.; von Albedyll, L.; Goessling, H. F.; Hendricks, S.; Juhls, B.; Spreen, G.; Willmes, S.; Belter, H. J.; Dethloff, K.; Haas, C.; Kaleschke, L.; Katlein, C.; Tian-Kunze, X.; Ricker, R.; Rostosky, P.; Rückert, J.; Singha, S.; Sokolova, J. MOSAiC drift expedition from October 2019 to July 2020: sea ice conditions from space and comparison with previous years. *Cryosphere* **2021**, *15*, 3897–3920.

2010, 16, 3154-3163.

(44) Krumpen, T.; Birrien, F.; Kauker, F.; Rackow, T.; von Albedyll, L.; Angelopoulos, M.; Belter, H. J.; Bessonov, V.; Damm, E.; Dethloff, K.; Haapala, J.; Haas, C.; Harris, C.; Hendricks, S.; Hoelemann, J.; Hoppmann, M.; Kaleschke, L.; Karcher, M.; Kolabutin, N.; Lei, R.; Lenz, J.; Morgenstern, A.; Nicolaus, M.; Nixdorf, U.; Petrovsky, T.; Rabe, B.; Rabenstein, L.; Rex, M.; Ricker, R.; Rohde, J.; Shimanchuk, E.; Singha, S.; Smolyanitsky, V.; Sokolov, V.; Stanton, T.; Timofeeva, A.; Tsamados, M.; Watkins, D. The MOSAiC ice floe: sediment-laden survivor from the Siberian shelf. *Cryosphere* **2020**, *14*, 2173–2187.

(45) Obbard, R. W.; Sadri, S.; Wong, Y. Q.; Khitun, A. A.; Baker, I.; Thompson, R. C. Global warming releases microplastic legacy frozen in Arctic Sea ice. *Earths Future* **2014**, *2*, 315.

(46) Krembs, C.; Eicken, H.; Deming, J. W. Exopolymer alteration of physical properties of sea ice and implications for ice habitability and biogeochemistry in a warmer Arctic. *Proc. Natl. Acad. Sci. U. S. A.* **2011**, *108*, 3653–3658.

(47) Gutow, L.; Eckerlebe, A.; Giménez, L.; Saborowski, R. Experimental Evaluation of Seaweeds as a Vector for Microplastics into Marine Food Webs. *Environ. Sci. Technol.* **2016**, *50*, 915–923.

(48) Bhattacharya, P.; Lin, S.; Turner, J. P.; Ke, P. C. Physical Adsorption of Charged Plastic Nanoparticles Affects Algal Photosynthesis. *J. Phys. Chem. C* **2010**, *114*, 16556–16561.

(49) Li, Y.; Liu, X.; Shinde, S.; Wang, J.; Zhang, P. Impacts of Microand Nanoplastics on Photosynthesis Activities of Photoautotrophs: A Mini-Review. *Front. Microbiol.* **2021**, *12*, 3324.

(50) Kanhai, L. D. K.; Gardfeldt, K.; Krumpen, T.; Thompson, R. C.; O'Connor, I. Microplastics in sea ice and seawater beneath ice floes from the Arctic Ocean. *Sci. Rep.* **2020**, *10*, 5004.

(51) Bergmann, M.; Wirzberger, V.; Krumpen, T.; Lorenz, C.; Primpke, S.; Tekman, M. B.; Gerdts, G. High Quantities of Microplastic in Arctic Deep-Sea Sediments from the HAUSGARTEN Observatory. *Environ. Sci. Technol.* **2017**, *51*, 11000–11010.

(52) Allen, D.; Allen, S.; Abbasi, S.; Baker, A.; Bergmann, M.; Brahney, J.; Butler, T.; Duce, R. A.; Eckhardt, S.; Evangeliou, N.; Jickells, T.; Kanakidou, M.; Kershaw, P.; Laj, P.; Levermore, J.; Li, D.; Liss, P.; Liu, K.; Mahowald, N.; Masque, P.; Materić, D.; Mayes, A. G.; McGinnity, P.; Osvath, I.; Prather, K. A.; Prospero, J. M.; Revell, L. E.; Sander, S. G.; Shim, W. J.; Slade, J.; Stein, A.; Tarasova, O.; Wright, S. Microplastics and nanoplastics in the marine-atmosphere environment. *Nat. Rev. Earth Environ.* **2022**, *3*, 393–405.

(53) Allen, S.; Allen, D.; Karbalaei, S.; Maselli, V.; Walker, T. R. Micro(nano)plastics sources, fate, and effects: What we know after ten years of research. *J. Hazard. Mater. Adv.* **2022**, *6*, No. 100057.

(54) Wiedmann, I.; Ershova, E.; Bluhm, B. A.; Nöthig, E.-M.; Gradinger, R. R.; Kosobokova, K.; Boetius, A. What Feeds the Benthos in the Arctic Basins? Assembling a Carbon Budget for the Deep Arctic Ocean. *Front. Mar. Sci.* **2020**, *7*, 224.

(55) Fernández-Méndez, M.; Wenzhöfer, F.; Peeken, I.; Sørensen, H. L.; Glud, R. N.; Boetius, A. Buoyancy Regulation and Fate of Ice Algal Aggregates in the Central Arctic Ocean. *PLoS One* **2014**, *9*, No. e107452.

(56) Stroeve, J.; Notz, D. Changing state of Arctic sea ice across all seasons. *Environ. Res. Lett.* **2018**, *13*, No. 103001.

(57) Tedesco, L.; Vichi, M.; Scoccimarro, E. Sea-ice algal phenology in a warmer Arctic. *Sci. Adv.* **2019**, *5*, No. eaav4830.

(58) Engel, A.; Piontek, J.; Metfies, K.; Endres, S.; Sprong, P.; Peeken, I.; Gäbler-Schwarz, S.; Nöthig, E.-M. Inter-annual variability of transparent exopolymer particles in the Arctic Ocean reveals high sensitivity to ecosystem changes. *Sci. Rep.* **2017**, *7*, 4129.

(59) Wollenburg, J. E.; Katlein, C.; Nehrke, G.; Nöthig, E. M.; Matthiessen, J.; Wolf- Gladrow, D. A.; Nikolopoulos, A.; Gázquez-Sanchez, F.; Rossmann, L.; Assmy, P.; Babin, M.; Bruyant, F.; Beaulieu, M.; Dybwad, C.; Peeken, I. Ballasting by cryogenic gypsum enhances carbon export in a *Phaeocystis* under-ice bloom. *Sci. Rep.* **2018**, *8*, 7703. (60) Søreide, J. E.; Leu, E. V. A.; Berge, J.; Graeve, M.; Falk-Petersen, S. Timing of blooms, algal food quality and *Calanus glacialis* reproduction and growth in a changing Arctic. *Glob. Chang. Biol.* (61) McMahon, K.; Ambrose, W. J.; Johnson, B.; Sun, M.; Lopez, G.; Clough, L.; Carroll, M. Benthic community response to ice algae and phytoplankton in Ny Ålesund, Svalbard. *Mar. Ecol. Prog. Ser.* **2006**, *310*, 1–14.

(62) Rantanen, M.; Karpechko, A. Y.; Lipponen, A.; Nordling, K.; Hyvärinen, O.; Ruosteenoja, K.; Vihma, T.; Laaksonen, A. The Arctic has warmed nearly four times faster than the globe since 1979. *Commun. Earth Environ.* **2022**, *3*, 168.

(63) Bergmann, M.; Almroth, B. C.; Brander, S. M.; Dey, T.; Green, D. S.; Gundogdu, S.; Krieger, A.; Wagner, M.; Walker, T. R. A global plastic treaty must cap production. *Science* **2022**, *376*, 469–470.