

1 **Predicted Vulnerability of Carbon in Permafrost Peatlands with Future Climate Change and**
2 **Permafrost Thaw in Western Canada**

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4 Treat, Claire C.^{1*}, Miriam C. Jones², Jay Alder³, A. Britta K. Sannel⁴, Philip Camill⁵, Steve Frolking⁶

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8 ¹ Alfred Wegener Institute Helmholtz Center for Polar and Marine Research, Telegrafenberg A45,
9 14473 Potsdam, Germany

10 ² Florence Bascom Geoscience Center, U.S. Geological Survey, Reston, VA USA

11 ³ Geosciences and Environmental Change Science Center, U.S. Geological Survey, Corvallis, OR USA

12 ⁴ Department of Physical Geography and Bolin Centre for Climate Research, Stockholm University,
13 Stockholm, Sweden

14 ⁵ Earth and Oceanographic Science Department and Environmental Studies Program, Bowdoin College,
15 Brunswick, ME USA

16 ⁶ Institute for the Study of Earth, Oceans, and Space, University of New Hampshire, Durham, NH USA

17

18 *Corresponding author email: Claire.treat@awi.de, phone: +49 (331)288-2136

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20 **Key points:**

- 21 • Simulated carbon balance in peatlands with and without permafrost using process-based model
- 22 from 8000 BP to 2100 CE
- 23 • Modeled decomposition losses from active layer peat (0.2 - 1.0 m) was the predominant carbon
- 24 source, not deeper peat or newly thawed permafrost
- 25 • Modeled new peat accumulation offset a large fraction of C losses but future changes in
- 26 vegetation productivity are poorly understood

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27 **Abstract**

28 Climate warming in high-latitude regions is thawing carbon-rich permafrost soils, which can release
29 carbon to the atmosphere and enhance climate warming. Using a coupled model of long-term peatland
30 dynamics (Holocene Peat Model, HPM-Arctic), we quantify the potential loss of carbon with future
31 climate warming for six sites with differing climates and permafrost histories in Northwestern Canada.
32 We compared the net carbon balance at 2100 CE resulting from new productivity and the decomposition
33 of active layer and newly-thawed permafrost peats under RCP8.5 as a high-end constraint. Modeled net
34 carbon losses ranged from -3.0 kg C m^{-2} (net loss) to $+0.1 \text{ kg C m}^{-2}$ (net gain) between 2015 to 2100.
35 Losses of newly thawed permafrost peat comprised 0.2 to 25% (median: 1.6%) of “old” C loss, which
36 were related to the residence time of peat in the active layer before being incorporated into the
37 permafrost, peat temperature, and presence of permafrost. The largest C loss was from the permafrost-
38 free site, not from permafrost sites. C losses were greatest from depths of 0.2 – 1.0 m. New C added to
39 the profile through net primary productivity between 2015-2100 offset ~40% to >100% of old C losses
40 across the sites. Differences between modeled active layer deepening and flooding following permafrost
41 thaw resulted in very small differences in net C loss by 2100, illustrating the important role of present-
42 day conditions and permafrost aggradation history in controlling net C loss.

43

44 **Plain language summary**

45 The thawing of permafrost in tundra, fen, bog, and other peatland wetlands can enhance climate change
46 through releasing carbon from the soil. Using a model for six sites in western Canada, we estimate how
47 much carbon will be lost by 2100 under a high-end emissions scenario. While these peatlands continue
48 to accumulate carbon in the surface soil layers through enhanced vegetation growth, more carbon is lost
49 from slightly deeper in the soil profile. As a result, most sites were projected to lose relatively small

50 amounts of carbon compared to how much they contain (< 5%). Little of the carbon lost was from newly
51 thawed permafrost, while the largest carbon losses were from the permafrost-free site. In sites where
52 permafrost thawed, the carbon losses were related to peatland and permafrost history.

53

54 **Introduction**

55 Peatlands in northern high latitudes store significant amounts of soil carbon (C), estimated at 450 to
56 1000 Pg C globally [Gorham, 1991; Nichols and Peteet, 2019]. In northern regions, an estimated 1.22 x
57 10⁶ km² of peatland contains permafrost, or 35% of northern peatlands [Hugelius *et al.*, 2020], which
58 protects peat C from decomposition and loss to the atmosphere when the peat is frozen [Tarnocai,
59 2006]. With warming temperatures in northern high latitudes associated with climate change, it is
60 important to understand whether peatlands (and permafrost peatlands) will continue be net C sinks or
61 whether some C stored in long-term peat C reservoirs will be released to the atmosphere and cause
62 feedback to the climate system [Frolking *et al.*, 2011; Schuur *et al.*, 2015]. While factors such as water
63 table position and temperature have long been known to control C exchange in peatlands [Gorham,
64 1991], permafrost peatlands may have some unique controls on C feedbacks. Once peat leaves the active
65 layer and becomes permafrost, any further decomposition is halted until it thaws. Some studies have
66 documented relatively large C losses post-thaw [Jones *et al.*, 2017; O'Donnell *et al.*, 2012], while others
67 have reported minimal losses [Estop-Aragonés *et al.*, 2018a; Heffernan *et al.*, 2020]. In permafrost
68 peatlands, decomposition of newly-thawed peat may be limited by the presence of highly degraded
69 material [Treat *et al.*, 2014] that has limited biogeochemical activity potential [Estop-Aragonés *et al.*,
70 2018a]. In general terms, the longer peat persists before becoming frozen into permafrost, the more
71 degraded the peat will be, and the less susceptible to decomposition upon subsequent permafrost thaw.

72 Empirical and experimental approaches to estimating losses of peatland C following permafrost
73 thaw have generated important hypotheses about the controls on C loss from peatlands following
74 permafrost thaw utilizing inter-site comparisons. Based on the analysis of peat types and decomposition
75 rates, and net C losses across a range of sites with different histories, the magnitude of net C loss is
76 hypothesized to be related to the relative timing of permafrost aggradation and peat deposition
77 [Heffernan *et al.*, 2020; Jones *et al.*, 2017; Treat *et al.*, 2014]. For example, chronosequences with
78 relatively large C losses accumulated peat which was quickly incorporated into permafrost over
79 millennia (e.g. syngenetic permafrost formation), resulting in minimally decomposed permafrost peat
80 that is then decomposed more readily upon thaw [Jones *et al.*, 2017; O'Donnell *et al.*, 2012; Treat *et al.*,
81 2014]. On the other hand, permafrost aggraded more recently in peatlands of Western Canada (Alberta,
82 Manitoba) than in some parts of Alaska [Treat and Jones, 2018; Zoltai and Vitt, 1990], suggesting that C
83 in permafrost peat was deposited substantially before incorporation into permafrost and therefore
84 subjected to greater decomposition. Results from empirical studies of peatlands in Alberta have shown
85 smaller C losses post thaw [Estop-Aragonés *et al.*, 2018a; Estop-Aragonés *et al.*, 2018b; Heffernan *et*
86 *al.*, 2020] than in Alaska [Jones *et al.*, 2017; Plaza *et al.*, 2019], offering some support for this
87 hypothesis.

88 While these losses of permafrost C are estimated using empirical modeling, the C losses from
89 permafrost are difficult to quantify directly using observations. Other *in-situ* processes, such as
90 enhanced net primary productivity in wetter, newly thawed peatlands, may mask any changes in peat C
91 exchange [Camill *et al.*, 2001; Prater *et al.*, 2007; Turetsky *et al.*, 2007]. Detecting the relatively small
92 fluxes of CO₂ or CH₄ from thawing permafrost peat or old, deep peat against the relatively large
93 magnitude of modern ecosystem respiration fluxes can be difficult using radiocarbon or δ¹³C isotopic
94 signatures [Estop-Aragonés *et al.*, 2018b], and the interpretation can be ambiguous because of the

95 fractionation associated with decomposition processes [*Dorrepaal et al.*, 2009; *Hicks Pries et al.*, 2015].
96 Furthermore, there is the complication of high interannual and spatial variability in CO₂ and CH₄
97 exchange that may mask long-term trends and affects conclusions drawn primarily from growing season
98 measurements [*Roessger et al.*, 2019; *Roulet et al.*, 2007; *Treat et al.*, 2018].

99 Interactions between temperature, carbon cycling in peatlands, and peatland hydrology are
100 difficult to untangle from field observations. Models can provide insights into how different factors
101 affect overall system behavior. Here, we compare the potential peat C losses with future warming and
102 permafrost thaw at six sites along a temperature and permafrost gradient in Canada, ranging from a
103 permafrost-free site in James Bay Lowland at the edge of discontinuous (sporadic) permafrost zone to
104 the sub-Arctic continuous permafrost zone [*Brown et al.*, 2002]. We use a process-based peatland
105 model, which allows us to examine how different factors, including predicted temperature increases, site
106 wetting and drying, productivity increases, and site permafrost history, combine and interact to affect the
107 peat C balance with warming and permafrost thaw. The goals of this study were to:

- 108 - Develop a model of long-term peatland processes and permafrost formation that can be
109 applied at multiple northern sites;
- 110 - Predict changes in peat C stocks in response to warming climate for a set of northern sites
111 that span a permafrost gradient;
- 112 - Determine what factors (e.g. mean annual air temperature, predicted temperature increase,
113 peat accumulation history, other factors) controlled the change in net C stocks.



114

115 **Figure 1.** Map of site locations included in this permafrost transect study across western Canada,
 116 ranging from permafrost-free in the south (JBL3), to discontinuous (Joey Lake, Selwyn Lake), and to
 117 continuous permafrost in the north (Ennadai Lake, Baillie Bog, Thelon-Kazan Peatland / TKP). Peat
 118 cores from each site were collected as part of individual studies (see Table 1) and analyses were
 119 included in an earlier synthesis project [*Treat et al.*, 2016]. The “Discontinuous” permafrost zone here
 120 also includes isolated and sporadic permafrost zones. Permafrost zone map source: Brown et al. [2002],
 121 treeline from *Olson et al.* [2001].

122

123 **Site Descriptions**

124 We selected six peatland sites that spanned the range of permafrost conditions in western Canada
125 where previous paleoecological studies have been completed (Figure 1, Table 1). The modern climate of
126 the study region is generally cold continental, with mean annual temperatures below or near 0 °C and
127 less than 750 mm of annual precipitation (Table 1). The peatland sites are located on the Canadian
128 shield, where peatlands formed following glacial retreat and the drainage of proglacial lakes [*Gorham et*
129 *al.*, 2007]. The southernmost site, JBL3, is a permafrost-free bog site in the James Bay Lowlands with
130 2.45 m of woody, herbaceous, and moss peat [*Holmquist and MacDonald*, 2014]. Joey Lake Peatland
131 (core JL2), was sampled from a 3 m deep permafrost bog located in a large peatland complex in the
132 isolated and sporadic permafrost zone near Thompson, Manitoba [*Camill et al.*, 2009]. Farther north in
133 discontinuous permafrost, a nearly 2 m deep peat plateau bog composed mainly of *Sphagnum* moss and
134 rootlet peat was sampled adjacent to Selwyn Lake [*Sannel and Kuhry*, 2008; 2009]. About 170 km
135 northeast in the continuous permafrost zone in the boreal-tundra ecotone, a 1.8 m core from a polygonal
136 peat plateau at Ennadai Lake also contained mainly *Sphagnum* moss and rootlet peat overlying fen peat
137 [*Sannel and Kuhry*, 2008; 2009]. Further north of treeline in continuous permafrost, Baillie Bog [*sic*]
138 (BB, core) was sampled from the middle of a high-centered polygonal [fen] peatland located in a
139 bedrock depression that contained ~2 m of Cyperaceae (sedge) peat with occasional depositions of
140 mineral material in the core [*Vardy et al.*, 2005; *Vardy et al.*, 2000]. Thelon-Kazan Peatland (TKP, core
141 TK1P2) was sampled from another high-centered polygon with nearly 2 m of peat that was mainly
142 composed of Cyperaceae with *Sphagnum* in the surface 35 cm and had a relatively high amount of
143 mineral material throughout the core [*Vardy et al.*, 2005; *Vardy et al.*, 2000]. Data available from cores
144 at all sites includes bulk density, carbon or organic matter content, radiocarbon dates, and either of plant
145 macrofossil analysis, a description of peat type, or an interpretation of peatland class. These data were
146 compiled in an earlier synthesis and are available in digital format for download [*Treat et al.*, 2016]

147 (doi: 10.1594/PANGAEA.863697). The dataset key (Variable: Auth.Site.CoreID) for the cores included
 148 in this study is as follows: JBL3 (HOL-JLB-03), Joey Lake (CAM-JL-2), Selwyn Lake (SAN-SEL-
 149 SL1), Ennadai Lake (SAN-ENL-1), Baillie Bog (VAR-BB-01), and TKP (VAR-TKP-01).

150

151

152 **Table 1.** Site names, locations, climatic information, peat height, basal ages, and core information for

153 the peat cores used in this study.

Site Name	Peatland type	Latitude (°N)	Longitude (°W)	Permafrost zone	Mean annual air temp. (°C)	Mean annual precip. (mm/y)	Peat height (cm)	Basal age (cal BP)	Fen-bog transition height (cm)
JBL3	Bog	52° 51.62'	89° 55.77'	None	0.5* ^a	728* ^a	245	7760	195
Joey Lake (core JLP2)	Permafrost bog	55° 27.95'	98° 9.80'	Isolated & Sporadic	-2.9* ^b	509* ^b	300	8000	176
Selwyn Lake (core SL1)	Peat plateau margin	59° 53'	104° 12'	Discontinuous	-3.3†	430†	197	6580	22
Ennadai Lake (core EL1)	Peat plateau margin	60° 50'	101° 33'	Continuous (~30 km south of tree-line)	-9†	290†	186	5810	49
Baillie Bog (BB, core BB1)	High-center polygon	64° 42.8'	105° 34.75'	Continuous (200 km N of tree-line)	-10.9* ^c	300 * ^c	197	7750	182‡
Thelon-Kazan Peatland (TKP, core TK1P2)	High-center polygon	66° 27.07'	104° 50.08'	Continuous (400 km N of tree-line)	-11.3* ^d	272* ^d	193	6620	183‡

154 * Climate normals from Environment Canada (1980-2010) for stations: ^aPickle Lake A, ^bThompson A,

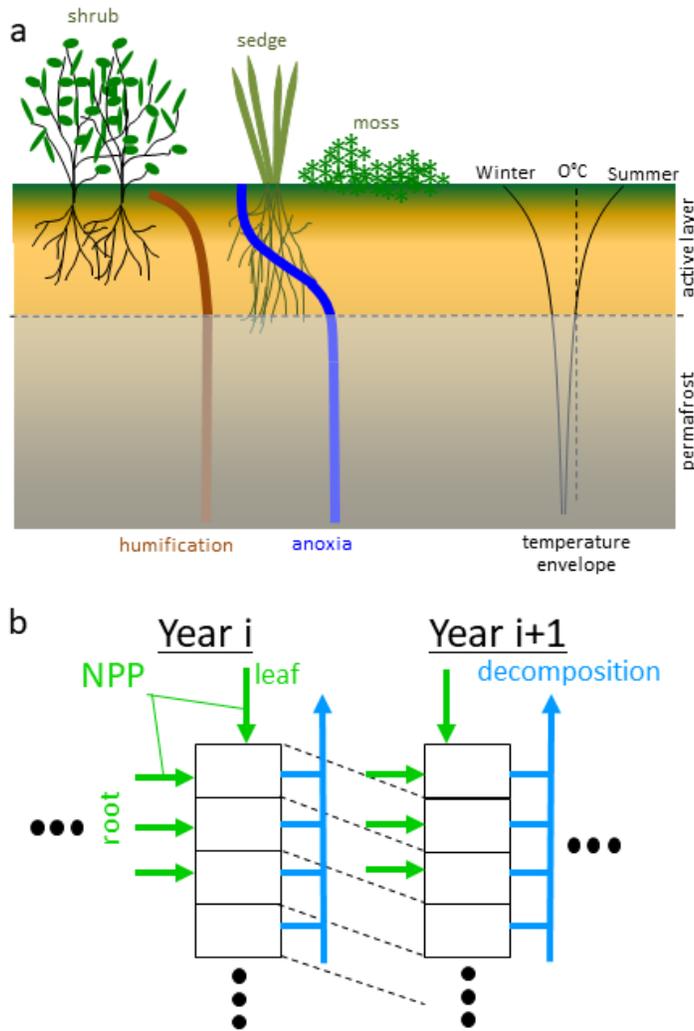
155 ^cLupin A, ^dBaker Lake A; https://climate.weather.gc.ca/climate_normals/; accessed 12 December 2019.

156 † Sannel & Kuhry 2008, 2009

157 ‡ No clear evidence of fen-bog transition. Here the authors describe a transition from herbaceous to

158 woody species in peat.

159



Basic structure of HPM. **(a)** In HPM, relative NPP of three plant functional types – moss, shrub, and sedge – varies with water table depth and active layer thickness. Anoxia profile is determined by depth below water table. Humification profile is determined by fresh root litter inputs and decomposition, as fraction mass remaining. Monthly temperature profiles computed by heat transfer with phase change, including a snowpack layer in winter. Permafrost occurs if $T_{\text{soil_layer}} < 0^{\circ}\text{C}$ for 24+ consecutive months. **(b)** All annual aboveground net primary productivity (NPP) is added as a surface litter cohort; belowground NPP is added to annual litter/peat cohorts based on rooting depth and profile (limited to active layer). Decomposition rates vary down the peat profile, controlled by litter quality and interacting profiles shown in panel (a). Annual peat cohorts accumulate through millennia of simulation to generate an age-depth profile characterized by relative amounts of moss, sedge, and shrub peat remaining.

160

161 **Figure 2.** Basic structure of HPM. (a) In HPM, relative NPP of three plant functional types – moss,
 162 shrub, and sedge – varies with water table depth and active layer thickness. Anoxia profile is determined
 163 by depth below water table. Humification profile is determined by fresh root litter inputs and
 164 decomposition, as fraction mass remaining. Monthly temperature profiles computed by heat transfer
 165 with phase change, including a snowpack layer in winter. Permafrost occurs if $T_{\text{soil_layer}} < 0^{\circ}\text{C}$ for 24+
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 169 interacting profiles shown in panel (a). Annual peat cohorts accumulate through millennia of simulation
 170 to generate an age-depth profile characterized by relative amounts of moss, sedge, and shrub peat
 171 remaining.
 172

173 **Methods**

174 *Model description*

175 We used HPM-Arctic to model peat formation since initiation to present and into the future. The
176 HPM-Arctic version integrates two earlier models: The Holocene Peat Model (HPM), a coupled carbon-
177 hydrologic model for peatlands [Frolking *et al.*, 2010], the Geophysical Institute Permafrost Lab soil
178 thermal model GIPL2 [Marchenko *et al.*, 2008]. Briefly, HPM simulates the development of a peat
179 profile over millennia, from initiation, using an annual litter cohort approach so that results can be
180 compared to dated peat cores (Figure 2). Rates of peat accumulation and decomposition are a function of
181 plant community composition (litter quality), modified by dynamic environmental conditions, including
182 water table level and water content in the unsaturated zone, and temperature profiles. Plant community
183 composition (i.e., relative litter inputs from different plant functional types) is dynamic, responding to
184 mean growing season water table depth and peat depth as a proxy for nutrient status. Annual net
185 primary productivity (NPP) is set equal to annual litterfall, the carbon input for peat accumulation. NPP
186 temperature sensitivity was modeled as a Q_{10} function, with a Q_{10} value of 1.8, based on an empirical
187 relationship between mean annual air temperatures and above-ground net primary productivity for
188 mosses, vascular plants, and trees that was developed for a transect of peatland sites in boreal Manitoba,
189 Canada [Camill *et al.*, 2001]. Peat bulk density in HPM is computed for each annual litter/peat cohort,
190 and increases non-linearly from a minimum to a maximum value ($50 - 130 \text{ kg m}^{-3}$) as the cohort loses
191 mass through decomposition [Frolking *et al.*, 2010]. The water table level is calculated from a simple
192 water balance model (precipitation minus evapotranspiration plus net run-on/run-off, and the net peat
193 water content determines the water table location, where the peat water content in the unsaturated zone
194 is a function of peat bulk density and distance above the water table [Frolking *et al.*, 2010].

195 HPM-Arctic has been modified from the original version of HPM in several ways. Principally, it
196 has been coupled to the Geophysical Institute Permafrost Lab soil thermal model GIPL2 [*Marchenko et*
197 *al.*, 2008], modified to include an accumulating peat layer on the soil surface [GIPL-2-peat; *Treat et al.*,
198 2013; *Wisser et al.*, 2011]. GIPL-2-peat is a soil thermal model that solves vertical soil heat transfer and
199 phase change using a numerical approximation accounting for soil type and soil water content; it is
200 driven by air temperature and includes a dynamic winter snowpack as a heat transfer layer. Soil
201 temperatures are calculated for a 100-m soil and bedrock column that has varying thermal properties
202 with depth and variable layer thicknesses, thinner at the surface (0.05 m minimum) and thicker down the
203 soil column into the bedrock (5 m maximum). As peat accumulates on the soil surface over centuries to
204 millennia (at rates generally $< 0.001 \text{ m yr}^{-1}$), the mineral soil and bedrock layers slowly descend deeper
205 into the simulated soil profile. Simulated soil temperatures are used to constrain rates of peat
206 decomposition, and variation in the active layer thickness is used instead of peat depth as an indication
207 of nutrient status, which impacts net primary productivity of vascular plants. Active layer thickness,
208 updated annually, is determined by identifying the soil thermal layer just above the top-most layer where
209 the temperature remains below 0° C for two years continuously, in accordance to the definition of
210 permafrost [*S A Harris et al.*, 1988].

211 In addition to soil temperature profiles, HPM-Arctic has (i) reduced the model time step from
212 annual to monthly for model drivers (air temperature and precipitation), peat profile water balance, and
213 decomposition (the soil thermal model operates at a daily timestep, using air temperatures interpolated
214 from the monthly values); (ii) reduced the number of plant functional types to three: moss, herbaceous
215 (including sedges and graminoids), and ligneous (woody species including shrubs), where for the
216 vascular plants, the inputs and decomposition of above and belowground litter are tracked separately;
217 and (iii) incorporated a simple ‘old-new’ carbon tracking algorithm, whereby after a specified year all

218 moss, sedge, and shrub plant litter gets labeled as 'new', so that its accumulation as peat and loss
219 through decomposition can be tracked separately from the older peat derived from plant litter inputs
220 prior to the specified year. This study used 2015 CE as both present-day and the boundary between new
221 and old carbon inputs. Model code for these runs and for HPM-Arctic is available for download (doi:
222 10.5281/zenodo.4647666).

223

224 **Model optimization and evaluation**

225 Some site-specific calibrations were done for several model parameters to capture variability (often not
226 reported) related to individual watershed and site characteristics (Table 2). Peat initiation often occurs in
227 a local topographic low, receiving run-on from the surrounding watershed; as the peat accumulates and
228 the peat surface rises, it can shift to a local topographic high point, and shed water (run-off) rather than
229 receive it [Charman, 2002]. The site-specific model parameters include the accumulating peat height at
230 which this shift from run-on to run-off occurred ($H_{\text{run-on/off}}$), and the peat height when initial fen-type
231 vegetation transitioned to bog-type vegetation (H_{FBT}). When the peat height exceeds the site-specific
232 H_{FBT} , the peatland transitions from a fen to a bog, which involves a decrease in annual NPP to a varying
233 degree [Rydin and Jeglum, 2006] modeled with a site specific fractional parameter ($F_{\text{NPP-bog}}$). With
234 greater lateral hydrological flow, and therefore a shorter water residence time in the saturated zone, fen
235 conditions are assigned a longer scale length (e-folding depth below the water table) to a full anoxia
236 impact on decomposition rate [e.g. Glaser *et al.*, 2016]. This is modeled with an anoxia scale length
237 parameter [Frolking *et al.*, 2010], which controls the decline in decomposition rate with depth below the
238 water table, emulating how far/quickly reduced electron acceptors are replenished below the water table.
239 This affects peat decomposition rates below the water table. If/when permafrost is present, frozen peat

240 decomposition rates are set to zero (also for seasonal winter frost), while decomposition persists in the
241 seasonally-thawed surface active layer.

242 Site-specific parameter values were determined from a combination of observations and/or
243 optimization routines (Table 2). The parameter H_{FBT} was determined by trial and error from the final
244 peat height and observations of the height of the fen-to-bog transitions (H_{FBT}) in the site core profiles;
245 the model parameter H_{FBT} generally was higher than in the observations. $H_{\text{run-on/off}}$ was determined by
246 trial and error from the agreement between the modeled peat height and observed peat height, as well as
247 the macrofossil composition, which indicated relative water table position over time (e.g. dry or wet).
248 The other three parameters ($F_{\text{NPP-bog}}$, and anoxia scale lengths for fen and bog) were determined from
249 minimizing the root mean squared error between the observed and modeled age-depth profiles, where
250 the age of the peat surface was assumed to be the year of sampling. For most sites, the optimization
251 routine was implemented using the *fmincon* solver in MATLAB. For sites with only two radiocarbon
252 dates (TKP, BB), the parameterization used was based on earlier runs at the next northernmost site
253 (Ennadai) and generalized parameters for temperate and boreal peatlands [Frolking *et al.*, 2010]. To
254 speed up the optimization process, the optimization routine used a simplified model version with a
255 constant time series of peat temperature profiles that was generated using the initial parameter values.
256 After solving for the optimal parameterization, we iterated this procedure by re-running the full model to
257 update the soil temperature profile using the optimized parameters, re-ran the optimization routine, and
258 re-ran the full model for a final time.

259 For calculating organic matter stocks and C stocks, model output (mass of peat) was multiplied
260 by mean values determined from a synthesis of over 10,000 peat layers spanning the permafrost region
261 [Treat *et al.*, 2016]. The conversion factor from peat to organic matter (OM) was $0.924 \text{ g OM g}^{-1} \text{ peat}$.
262 The conversion factor from organic matter to carbon was $0.495 \text{ g C g}^{-1} \text{ OM}$.

263 We calculated the residence time of peat in the active layer before being incorporated into
 264 permafrost. The active layer residence time was dynamic over time, as both peat height and active layer
 265 thickness change every year, so this value was determined using the 500-year litter cohorts tracked in
 266 HPM. HPM-Arctic tracks several metrics for these 500-year litter cohort markers every year, including
 267 the depth below the peat surface and height above the mineral soil surface. We found the height of the
 268 permafrost in the peat profile in each year and calculated the year when each litter cohort was first
 269 incorporated into the permafrost. Knowing the age of the 500-year litter cohort when it entered the
 270 permafrost allowed us to calculate the residence time of that peat cohort in the active layer.

271 **Table 2.** Site-specific parameterization used in HPM-Arctic. * Indicates that parameter value was
 272 determined using an optimization routine (Optimization methods).
 273

Parameter	Description	JBL3	Joey Lake	Selwyn Lake	Ennadai Lake	Baillie Bog (BB)	Thelon-Kazan (TKP)
H _{run-on/off}	Height run-on/off (m)	1.50	0.31	0.40	0.75	2.2	4.5 ^a
H _{FBT}	Height of fen-bog transition (m)	2.0	2.2	0.6	2.2	2.3	2.1
F _{NPP-bog}	NPP multiplier at H _{FBT}	0.73*	0.55	0.55	0.48*	0.48	0.48
AnoxiaScale-Fen	anoxia scale length in fen when height < H _{FBT} (m)	1.10*	1.11*	2.46*	2.05*	3.0	3.0
AnoxiaScale-Bog	anoxia scale length in bog when height > H _{FBT} (m)	0.81*	0.28*	1.075*	0.21*	1.2	1.2
Max NPP	Maximum annual NPP under ideal conditions (kg m ⁻² y ⁻¹).	1.5	1.5	1.5	1.5	1.1	1.1
Ndates	Number of ¹⁴ C dates in profile*	6	11	13	4	2	2
RMSE	Root mean square error, model vs. observations	26.4	182.1	70.9	34.1	27.1	21.3

274 * The age of the peat surface was assumed to be the same as the year of sampling and included in the age-depth model.

275 ^a A value deeper than the peat height indicates that the peat will continue to receive run-off until the threshold is
 276 reached.

277

278 Model scenarios

279 We explored two possible water table scenarios resulting from warmer soil temperatures and
 280 permafrost thaw: perched and flooded. The perched water table scenario reflected a water table perched

281 on top of the permafrost, which results in a drier soil profile as water drains laterally due to differences
282 in elevation. This feature is common in peat plateaus and palsas [Zoltai, 1993], which are elevated from
283 the surrounding peatlands from uplift associated with permafrost formation [Seppälä, 2011], resulting in
284 water drainage to adjacent lower-lying areas. Because the water table is perched on top of the
285 permafrost, as the depth to permafrost increases with warming soil, the depth to the water table also
286 increases and sites become drier [Haynes *et al.*, 2018; Osterkamp *et al.*, 2009]. No changes were needed
287 to the model setup to reproduce this behavior.

288 The flooded scenario represents a peatland with thawing permafrost that is receiving water from
289 adjacent uplifted areas [Osterkamp *et al.*, 2009; Zoltai, 1993]. This would be fairly analogous to a
290 thermokarst peatland, where the peat surface becomes flooded due to permafrost thaw, ice melt, and
291 resulting subsidence of the peat surface [Osterkamp *et al.*, 2000]. However, while HPM-Arctic simulates
292 frozen peat in the peat profile, it does not simulate the formation or degradation of ice lenses, and it does
293 not account for volume or peat height changes associated with ice formation, accumulation, and thaw
294 [e.g. Seppälä, 2011]. Instead, we used the flooding scenario to mimic the lateral redistribution of water
295 from the surrounding peat associated with permafrost thaw and collapse [e.g. Osterkamp *et al.*, 2009;
296 Seppälä, 2011] by resetting the $H_{\text{run-on/off}}$ parameter. The new value for $H_{\text{run-on/off}}$ was set to the peat
297 height at the time when two conditions were met: (i) permafrost thaw had reached a mean active layer
298 thickness >1.25 m for the previous 10-year period (active layer thickness greater than 1.25 is not
299 measurable with a typical frost probe and not likely to be detected without intensive GPR surveys), and
300 (ii) peat height decreased (net peat added was less than peat lost over the previous 10-year period).
301 Resetting $H_{\text{run-on/off}}$ to the new peat height had the effect of wetting the peat and bringing the water table
302 near the surface, as is commonly observed in thermokarst features [Osterkamp *et al.*, 2000]. Because
303 the flooding scenario was determined dynamically within the model, not all sites experienced

304 thermokarst flooding, either because the active layer thickness did not exceed 1.25 m or because peat
305 was not lost.

306

307 *Model climate drivers*

308 We utilize the TraCE-21ka transient simulations [*Liu et al.*, 2009 <https://www.earthsystemgrid.org/project/trace.html>] to drive HPM-Arctic with monthly temperature and precipitation
309 climate forcings from 8000 B.P. to 1990 CE (8 Kyr time series). The TraCE-21ka simulations are driven
310 by paleo changes in greenhouse gases, insolation, and paleogeography as sea level rises from the
311 melting of the large northern hemisphere ice sheets. Monthly temperature and precipitation time series
312 were extracted from the coarse $3.75^\circ \times \sim 3.75^\circ$ grid by bilinear interpolation to the six peatland site
313 locations (Figure 1, Table 1). The simple “delta method” bias correction was applied to the TraCE-21ka
314 output by converting the 8 Kyr time series to anomalies relative to 1950-1990 CE and applying the
315 anomalies to a modern observed gridded data set [CRU TS v3.32; *I Harris et al.*, 2014]. TraCE-21ka
316 temperature anomalies are applied additively to modern observations, whereas precipitation anomalies
317 are applied as scalars. To continue the 8 Kyr timeseries into the future, we adopt the CCSM4 RCP8.5
318 simulation from the Coupled Model Intercomparison Project Phase 5 [CMIP5; *Taylor et al.*, 2012]. The
319 RCP8.5 scenario was chosen as an end-member to bracket the greatest projected changes in temperature
320 and precipitation. CCSM4 [*Gent et al.*, 2011] is the successor to the CCSM3 model used in the TraCE-
321 21ka simulations, which we chose for consistency. While focusing on one model projection is a
322 limitation, CCSM4 has an equilibrium climate sensitivity (the response to a doubling of atmospheric
323 CO₂) of 2.9 °C, which is similar to the CMIP5 ensemble multi-model mean of 3.2 °C [*Flato et al.*,
324 2013], giving us confidence the projected temperature changes at our sites are not unreasonable relative
325 to the full CMIP5 ensemble. To extract CCSM4 projection time series at the peatland site locations, we
326

327 use the same process of bilinearly interpolating from the CCSM4 grid (0.9° x 1.25°) and applying CRU
328 bias correction using the 1950-1990 CE climatology period. In our analysis and results below, HPM-
329 Arctic is driven by TraCE-21ka output prior to 1990 CE and CCSM4 afterwards. The RCP8.5
330 simulations spans 2005 through 2100 CE.

331

332 *Statistical analysis and data analysis*

333 We used two basic statistical analyses in our evaluation of model results. We tested for differences
334 between present day and future C stocks in the two scenarios (perched, flooded) across the sites using a
335 t-test (R command: t.test). We used a paired t-test when comparing the differences between time periods
336 (present-day, 2100 CE) to account for differences between the sites and scenarios. We tested several
337 hypotheses for controls on total C loss in the future scenarios, including climatic conditions and changes
338 (present-day observed mean annual temperatures and precipitation, projected air temperatures and
339 precipitation from 2071-2100 CE, and the change between the two time periods), and some peat
340 characteristics, including basal ages, peat C stocks in 2015 CE, and mean degree of peat decomposition
341 across the peat profiles in 2015 CE. We used linear regressions between the changes in C stocks by 2100
342 CE and the predictor variables to test for significance of the predictors (R command: lm). All statistical
343 analyses were conducted using R statistical software [R Core Development *Team*, 2008].

344 In order to compare modeled C losses with observations, we used the peat C stock in 2015 CE,
345 and the results from old/new C tracker in 2100 CE. Net C loss (or gain) was calculated from the
346 difference in total C stocks in 2015 and 2100 CE; mean annual rates were calculated by dividing the
347 difference by Δt (85 years) and multiplying by 100 years century⁻¹ for century rates.

348

349

350 **Results and discussion**

351 *Model evaluation*

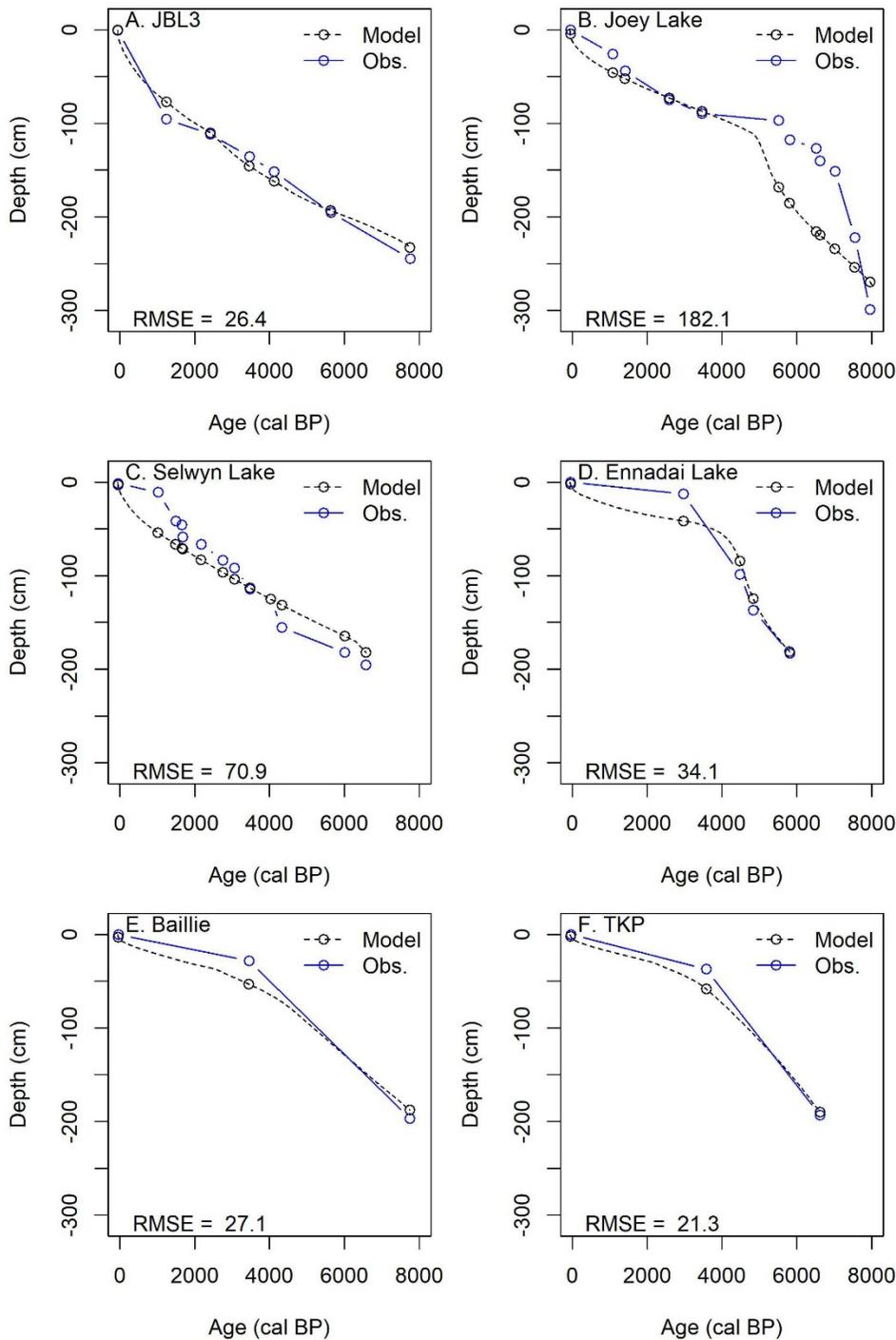
352 At each site, the peat profiles simulated by HPM-Arctic were compared against peat core observations,
353 including age-depth profiles. HPM-Arctic was able to reproduce the patterns of peat accumulation (i.e.,
354 peat age-depth profiles) at all sites using site-specific parameters related to hydrology and vegetation
355 productivity (Figure 3, Table 2). The good agreement between the model and observations indicates that
356 the model is capable of simulating realistic rates of peat accumulation and total peat height for sites
357 across the northern permafrost region (Figure 3). The final peat height in present-day did not differ
358 significantly between observed cores and the modeled profiles ($t=-0.52$, d.f.=9.6, $P=0.61$). The
359 simulated organic matter density profile in the peat was in broad agreement with data from the peat
360 cores across the sites, however HPM-Arctic profiles were always much smoother than observation
361 (Figure S1). In particular, the model underestimated organic matter density near the peat surface (to
362 about 20-40 cm depth) at several of the most northern sites (e.g., Ennadai, Baillie Bog, and TKP), and
363 overestimated shallow peat bulk density at the southern site (JBL3). Mean modeled peat C stocks across
364 the six sites did not differ significantly from the observational mean (mean modeled = 104 kg C m^{-2} ;
365 mean observed = 110 kg C m^{-2} ; $t=-0.41$, d.f.=9.5, $P= 0.69$). The model had only partial success in
366 simulating the dominant PFT composition of the peat profiles, and while the simulations generated peat
367 that was a mix of all PFTs, the modeled profiles were predominantly moss (Table 3).

368

369 **Table 3.** Field observations and model results for present day (2015: means of 2006-2015 CE, control) and future scenarios (2100:
370 means of 2091-2100 CE) for peat height, peat C stocks, water table level, maximum annual active layer thickness, and dominant
371 vegetation types, both by productivity (e.g. mean decadal NPP of surface vegetation) and preserved in the peat core record.

Site	Year / scenario	Scenario	Peat height (m)	Peat C (kg C m ⁻²)	Water table level (m below surf)	Active layer thickness (m)	Dominant vegetation – NPP	Dominant vegetation – Peat
JBL3	2008	Observed	2.44	105.0	-	NA		Woody/moss
	2015 – C	Control	2.30	120.2	0.07	NA	Sedge	Sedge
	2100 – P	Perched	2.22	117.3	0.20	NA	Shrub	Moss
Joey Lake	2001	Observed	3.00	154.0	(unknown)	(unknown)	-	-
	2015 – C	Control	2.72	138.3	0.20	2.22	Shrub	Moss
	2100 – P	Perched	2.63	135.6	0.35	> 2.63	Shrub	Moss
	2100 – F	Flooded	2.66	136.7	0.23	> 2.66	Shrub	Moss
Selwyn Lake	1993	Observed	1.97	84.9	(dry)	0.47	Shrub	Woody/moss
	2015 – C	Control	1.82	91.5	0.19	0.62	Shrub	Moss
	2100 – P	Perched	1.82	91.4	0.34	> 1.82	Shrub	Moss
	2100 – F	Flooded	1.82	91.6	0.24	> 1.82	Shrub	Moss
Ennadai Lake	2002	Observed	1.86	74.3	(dry)	0.41	Shrub	Woody/moss
	2015 – C	Control	1.81	83.5	0.27	0.73	Shrub	Moss
	2100 – P	Perched	1.78	82.4	0.40	0.87	Shrub	Moss
	2100 – F	Flooded	1.78	82.4	0.40	0.87	Shrub	Moss
Baillie Bog	1993/4	Observed	1.97	122.3	(0.40)	0.40	Shrub	Sedge/Shrub
	2015 – C	Control	1.88	96.8	0.06	0.85	Sedge	Moss
	2100 – P	Perched	1.85	95.6	0.21	0.66	Shrub	Moss
TKP	1993/4	Observed	1.93	107.5	(0.40)	0.40	Shrub	Sedge/Shrub
	2015 – C	Control	1.90	88.6	0.03	0.87	Sedge	Moss
	2100 – P	Perched	1.85	87.0	0.15	0.79	Sedge	Moss

372



374

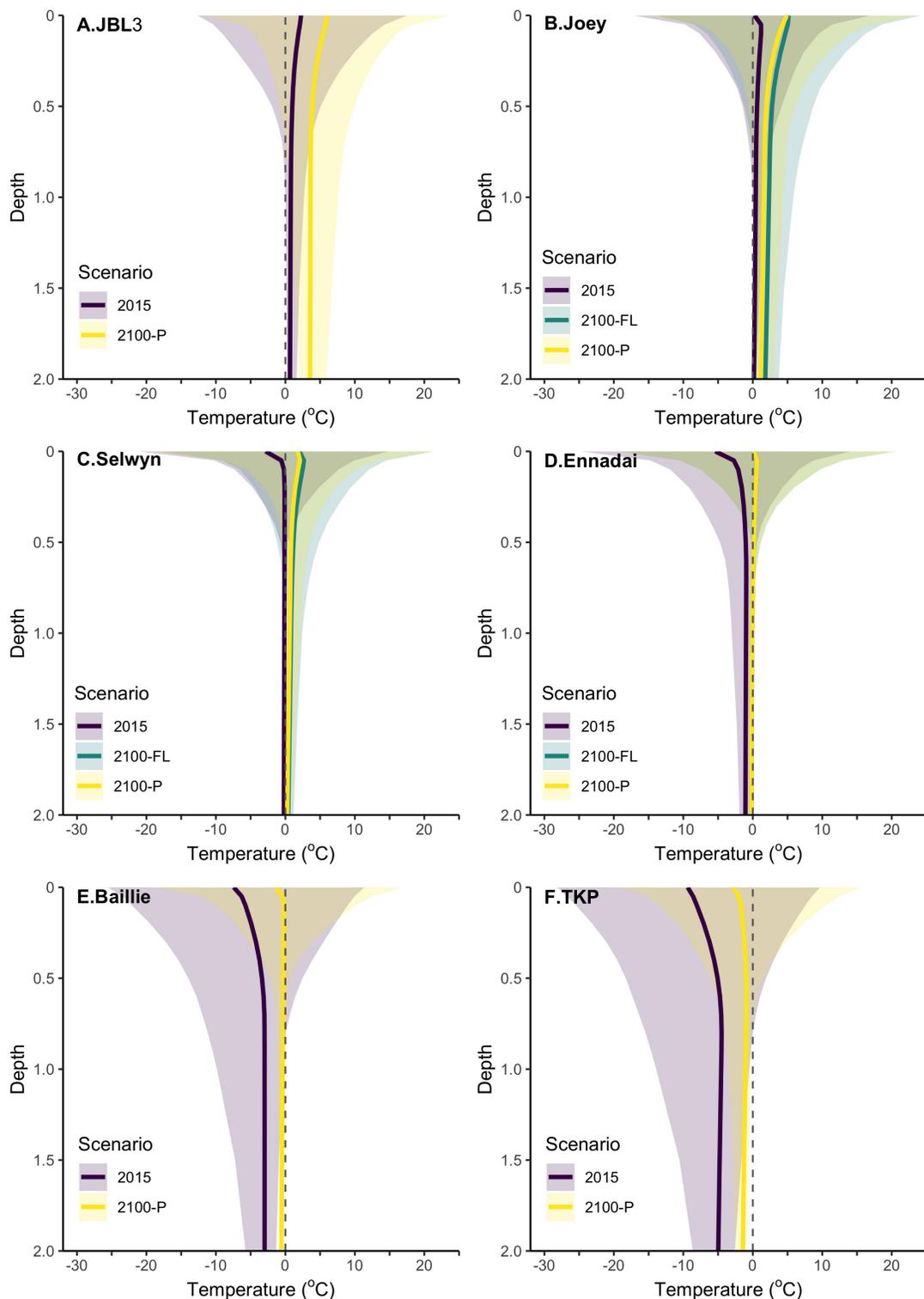
375 **Figure 3.** Model and observed age-depth profiles for cores in study, listed from south to north: A) JBL3,

376 B) Joey Lake, C) Selwyn Lake, D) Ennadai Lake, E) Baillie Bog, F) TKP-Wet. RMSE represents root

377 mean squared error of the difference between modeled and observed depths at the known radiocarbon

378 age sampling points from the observations. Note that model age-depth profiles are continuous results,

379 with labels added at depths of age observations in cores for comparison.



380
 381 **Figure 4.** Distribution of modeled peat temperatures for 2006-2015 and 2091-2100 in A) JBL3, B) Joey
 382 Lake, C) Selwyn Lake, D) Ennadai Lake, E) Baillie Bog, and F) TKP. Heavy lines represent the mean
 383 annual peat temperatures, while the shaded areas represent the temperature range between the mean
 384 minimum annual and mean maximum annual peat temperature over the periods of interest. 'FL' and 'P'
 385 refer to flooded and perched water table scenarios (see text).

386

387 *Model permafrost simulations and future projections*

388 HPM-Arctic successfully simulated permafrost in the sites with permafrost and no permafrost in
389 the permafrost-free site in the present day (Table 3). At JBL3, the site without permafrost, the surface
390 peat experienced seasonal freezing but the deeper peats remained thawed throughout the year (Figure 4).
391 In the permafrost sites, modeled active layer thickness ranged from 0.6 – 2.2 m (Table 3), which was
392 generally deeper than observations (Table 3). Across the sites with permafrost, we simulated a mean of
393 $47 \pm 35 \text{ kg C m}^{-2}$ (sd) or 44% of total C stocks in the active layer, while a mean of $52 \pm 15 \text{ kg C m}^{-2}$ or
394 56% was contained within the permafrost.

395 By 2100 CE, the predicted mean annual air temperatures under RCP8.5 reached above freezing
396 at the southern three sites and were substantially warmer at the northern sites (Figure S2, Table S1). The
397 projected warming increased the active layer thickness in several sites with permafrost after 2000 CE
398 (Table 3), resulting in the complete thaw of permafrost within the peat profile at the two southern
399 permafrost sites, Joey and Selwyn Lake (Figure 4), while active layer thickness increased at Ennadai
400 after 2050 (Figure S3). Permafrost remained in the peat profile for the three northernmost sites (Figure
401 4), with projected active layer thickness becoming shallower at Baillie Bog and staying similar at TKP
402 (Table 3; Figure S3). Projected precipitation changes by 2100 generally fell within the decadal ranges of
403 modern precipitation amounts (Figure S4) but did increase by 2091-2100 relative to the present CRU
404 data (Table S1). The projected warming increased modeled NPP across all sites after 2000 (Figure S5).
405 In all sites, simulated NPP increased during the 21st century (Figure S5), ranging from 25% (Joey Lake)
406 to nearly tripling (Selwyn Lake). At all sites, this was dominated by increases in woody (shrub) NPP,
407 which increased by a mean of >200% across the sites.

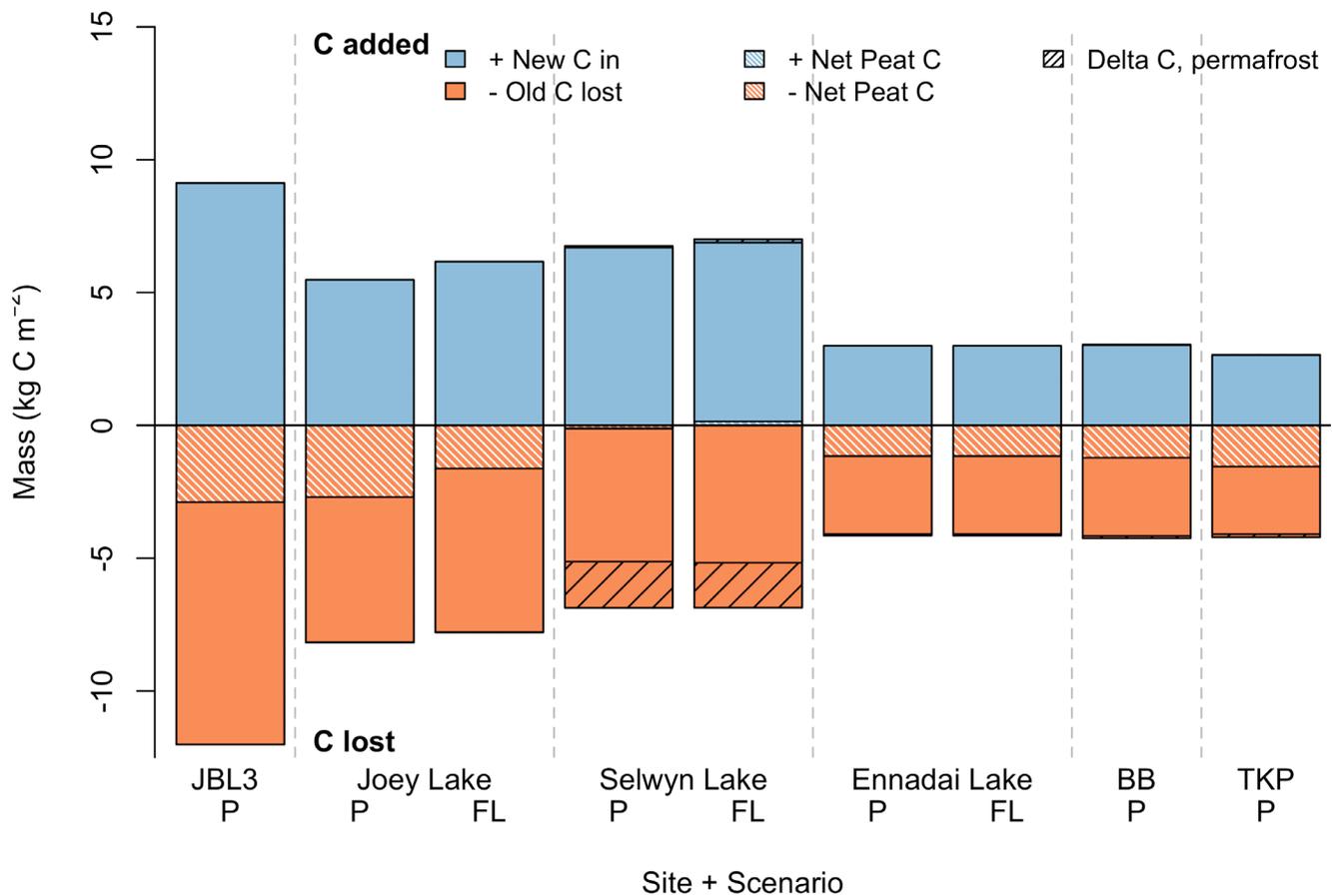
408 HPM-Arctic simulated reduced net peat C stocks across all scenarios and sites by 2100 CE in
409 response to 21st century warming (Figure 5). Across all sites, HPM predicted a loss of old peat
410 accumulated before 2015 CE, ranging from -13.2 to -4.1 kg C m⁻² between 2015 CE and 2100 CE. New
411 peat was added from the additions of net primary productivity (Figure 5), which offset between 40% and
412 100% of C losses. Consequently, the net loss across the sites was predicted to be between -3.0 and +0.1
413 kg C m⁻² by 2100 CE (Figure 5), with a mean loss of -1.6 kg C m⁻². This was a statistically significant
414 decrease relative to C stocks in 2015 CE ($t=-4.5$, d.f.=9, $P= 0.002$; Table 3), but < 5% of the modeled
415 peat C stocks at these sites. Median projected C loss from permafrost was -0.1 kg C m⁻² by 2100 CE,
416 and ranged from -1.7 kg C m⁻² to -0.01 kg C m⁻² (Figure 5). We hypothesized that the magnitude of net
417 peat C loss throughout the profile in 2100 CE could be predicted by the climate drivers (temperature,
418 precipitation), changes in the climate drivers, and the characteristics of the peat in 2015 CE. However,
419 none of these factors were significant predictors of the magnitude of C stock change by 2100 CE ($P >$
420 0.05).

421 The model allowed us to examine the depth-distribution of C losses and gains in the peat profiles
422 across the sites between 2015 CE and 2100 CE. Net C additions occurred in the surface 20 cm of peat in
423 the three southernmost sites (JBL3, Joey Lake, Selwyn Lake, mean: $+2.7 \pm 1.4$ kg C m⁻²), while losses
424 of old C from the surface 20 cm were roughly equivalent to new C inputs in the northern sites (Ennadai,
425 Baillie Bog, and TKP, mean: -0.2 ± 0.3 kg C m⁻²) resulting in a near-neutral C balance from the surface
426 20 cm of peat (Figure 6A,B). Across all sites, net C losses were greatest from 20-50 cm depths despite
427 small additions of new C (Figure 6A,B); net losses ranged from -3.8 to -0.7 kg C m⁻² and comprised 30-
428 50% of losses of old C from the profile. Net C losses from 0.50-1.0 m depths were smaller (mean: $-0.8 \pm$
429 0.6 kg C m⁻²; range: -2.0 to -0.2 kg C m⁻²) (Figure 6A,B). Losses of C from deep peats (> 1m depth)

430 were generally very small (-0.2 to -0.0 kg C m⁻²) with the exception of deep peat losses occurring in
431 JBL3 and Selwyn Lake, which ranged from -0.8 to -0.7 kg C m⁻² by 2100 CE (Figure 6A,B).

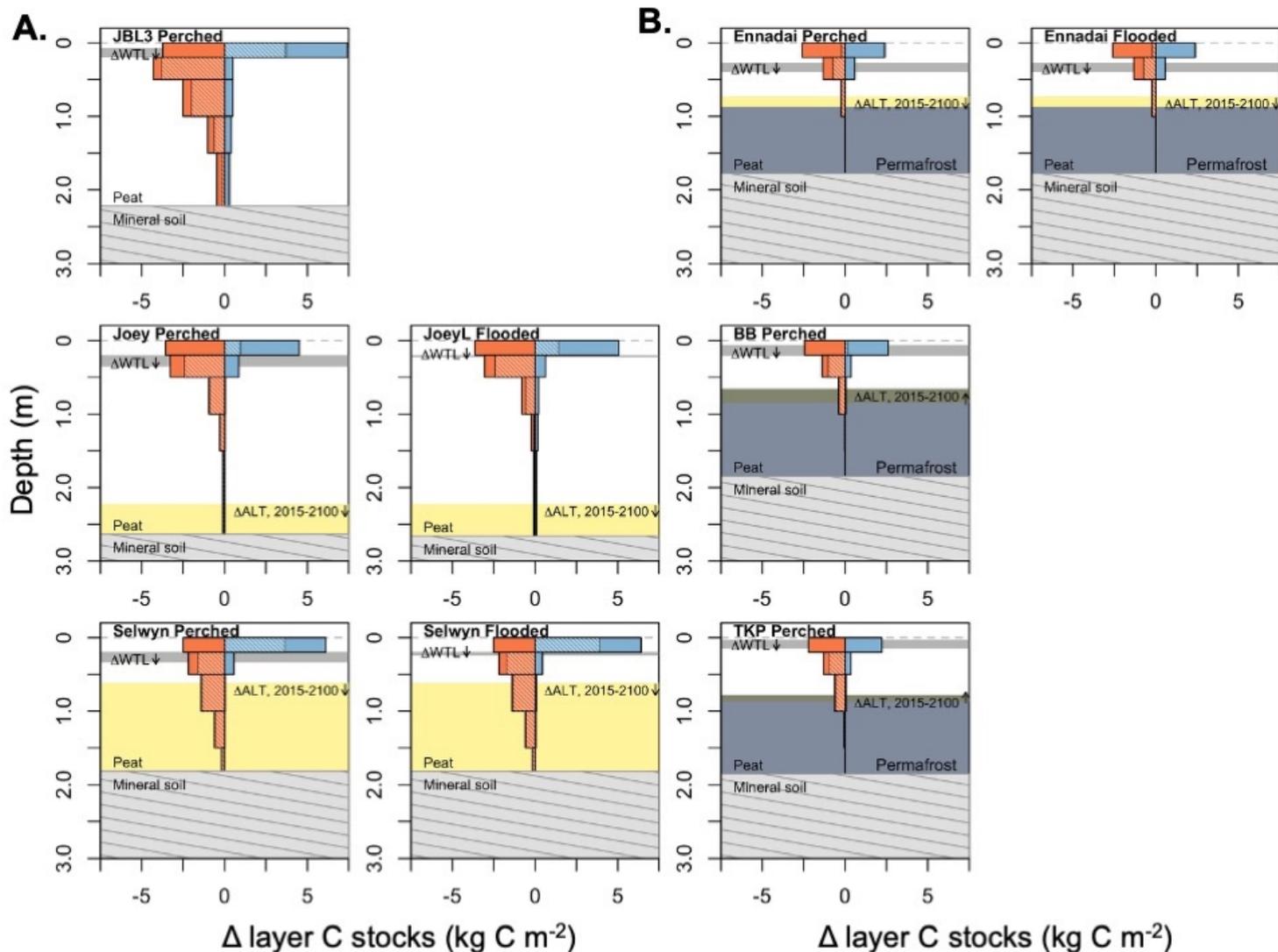
432 The two water table scenarios altered the hydrology towards drier conditions in the “perched”
433 water table scenario and wetter conditions in the “flooded” water table scenario in the sites that
434 experienced active layer deepening (Table 3). In the “perched” water table scenarios, the deepening of
435 the active layer at Joey Lake, Selwyn Lake, and Ennadai Lake resulted in ~ 14 cm decrease in the water
436 table (Table 3), which was essentially perched (constrained) atop the permafrost (Figure S3). While the
437 net primary productivity of the herbaceous plant functional types and the total NPP were lower in the
438 perched scenario than the flooded scenario (Figure S5), the thaw scenarios had little effect on the net C
439 stocks (Figure 5). Compared with the “flooded” scenario, where the water table decreased by only 7 cm
440 (Table 3), net peat C losses from the perched scenario were as much as -1.4 kg C m⁻² larger than in the
441 flooded scenarios by 2100 CE (Figure 5), but were not significantly different between the two scenarios
442 ($t=-2.1$, d.f. = 3, $P = 0.13$). A comparison of net C change across peat depths for the two water table
443 scenarios for Joey, Selwyn, and Ennadai Lake shows similar trends with depth across sites and only
444 small differences in magnitudes of C losses and gains between the two scenarios (Figure 6A,B).

445



446

447 **Figure 5.** Change of modeled peat C stocks at the sites along the permafrost gradient between 2015 and
 448 2100 under RCP8.5. “New” (blue) indicates the net C added to the peat through net primary productivity
 449 and lost from decomposition between 2015 and 2100, while “Old” (orange) indicates the total C fixed
 450 prior to 2015 that was lost between 2015 and 2100. Diagonal white shading indicates the magnitude of
 451 net change in peat C stocks due to old C losses and new C gains. Black diagonal lines indicate changes
 452 in permafrost C either from additions (black on blue) or losses (black on orange). Scenarios for all sites
 453 include perched water table (“P”), where the water table remains perched on top of the permafrost or
 454 mineral soil; flooded (“FL”), where the peat surface becomes flooded from localized runoff as
 455 permafrost thaws (see text; not triggered at all sites).



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Figure 6. Change in peat C stocks between 2015 CE and 2100 CE by depth for the study sites. A) Southern sites including JBL3 (top), Joey Lake (middle), Selwyn (bottom) with perched water table scenarios in the left column and flooded (thermokarst) scenarios in the right column; B) northern sites including Ennadai (top), Baillie Bog (middle) and TKP (bottom). Losses of “old” C fixed before 2015 is shown in orange, net new added C after 2015 is in blue, the net of old C losses and net new C gains is shown by lighter orange and blue shaded areas. Changes in water table levels are indicated by horizontal gray bars and Δ WTL, changes in active layer thickness (e.g. permafrost thaw) are indicated by yellow zones and Δ ALT, both for 2091-2100 relative to 2006-2015. Permafrost peat in 2100 CE is shown in blue-gray.

463 **Discussion**

464 *Peatland and permafrost history as a driver of C loss post-thaw*

465 A few empirical studies have quantified the potential for soil C loss following permafrost thaw in
466 organic soils by combining site-level observational approaches with empirical modeling. A recent study
467 based on repeated measurements projected (old) soil C losses ranging from -25 to -20 kg C m⁻² century⁻¹
468 by 2100 for a tundra site in Alaska experiencing permafrost thaw [Plaza *et al.*, 2019]; HPM-Arctic
469 simulated smaller losses of old C, ranging from -15.5 to -4.8 kg C m⁻² century⁻¹. Using a
470 chronosequence of thawed permafrost peatlands to account for the additions of new C to the peat profile
471 as well as losses of old peat from deeper in the peat profile, Jones *et al.* [2017] estimated *net* loss rates
472 of peat C from following thaw of -35 to -5 kg C m⁻² century⁻¹ from boreal sites in Alaska. HPM-Arctic
473 modeled range for *net* C loss rates (including net new C added and old C lost) across all permafrost sites
474 in this study was -15.5 to + 0.2 kg C m⁻² century⁻¹. These net C loss rates agree well with recent
475 observations following thaw from a boreal permafrost peatland in Canada, which range from -10.6 kg C
476 m⁻² century⁻¹ to +2.7 kg C m⁻² century⁻¹ [Heffernan *et al.*, 2020]. There is substantial disagreement
477 among the different empirical results from these different sites, which may be due simply to site
478 differences, but multi-site comparisons have shown underlying differences in substrate related to
479 permafrost and peatland histories. The timing of permafrost aggradation relative to peat formation is
480 hypothesized to be a major control on potential C loss from peatlands due to how decomposed the peat
481 substrate is at the time when it is frozen [Jones *et al.*, 2017; Treat *et al.*, 2014].

482 HPM can be useful to investigate the hypothesis that the relative timing of peat deposition to
483 permafrost aggradation controls C loss post-thaw. We quantified the length of time between peat
484 deposition and the peat entering the permafrost in HPM, i.e., its residence time in the active layer and
485 length of time the peat is subject to decomposition. In this permafrost transect study, only the two

486 southern permafrost sites were projected to have meaningful permafrost thaw by 2100 CE, which would
487 expose substantial previously frozen peat to decomposition. The model showed that permafrost thawed
488 completely within the peat profile at these sites by 2100 CE (Figure 6A, Table 3) and the peat was
489 subsequently vulnerable to decomposition. However, simulated net C lost from permafrost at Joey Lake
490 was $\sim 0 \text{ g C m}^{-2}$ and 1.7 kg C m^{-2} at Selwyn by 2100 CE (Figure 5), while deeper peat losses from
491 depths of 0.5 m to 1.5 m were 33-66% smaller at Joey Lake (-0.7 to -1.2 kg C m^{-2}) than at Selwyn Lake
492 (-1.9 to -2.0 kg C m^{-2} ; Figure 6A) despite warmer peat temperatures (Figure 4B,C). The modeled peat at
493 Joey Lake in the 0.5 m to 1.5 m depths was slightly more decomposed than at Selwyn Lake (median:
494 89% mass lost vs. 87% mass lost), but a better explanation of the difference could be the length of the
495 residence time that the peat was in the active layer and vulnerable to decomposition before being frozen
496 into the permafrost. At Joey Lake, the peat between 0.5 m and 1.5 m was in the active layer for ~ 2250
497 years prior to being incorporated into permafrost, whereas at Selwyn Lake, the peat was incorporated
498 into the permafrost after only 1200 years, limiting the length of time for decomposition in the past and
499 increasing the vulnerability of permafrost. Thus, the amount of time peat resides in the active layer or
500 prior to permafrost aggradation can influence the degree decomposition prior to incorporation into the
501 permafrost and can subsequently dictate the amount of post-thaw carbon loss. Therefore, understanding
502 the site history can be very important for projecting potential C loss with permafrost thaw.

503

504 *Peat C vulnerability to warmer temperatures: Permafrost thawing acts as a buffer to warming*

505 Site history is not the only important predictor of potential C loss in the future with climate
506 warming and permafrost thaw. We initially hypothesized that projected peat C loss by 2100 CE would
507 be greatest from sites that experienced a shift in mean annual air temperature from below freezing to
508 above 0°C (e.g. Joey Lake and Selwyn Lake, Figure S2) because of the resulting permafrost thaw. While

509 the magnitudes of net C losses were not significantly correlated with changes in environmental variables
510 or peat characteristics, changes in peat temperature and moisture were the strongest at JBL3 (Figure 4,
511 Table 3) despite having a smaller increase in air temperature than at more northern sites (+3.8 °C at
512 JBL3 vs. +6.5 °C at Joey Lake; Table S1). At JBL3, warmer peat temperatures from 0.5 to 2 m depth
513 (with a mean annual temperature of +3°C (range 0-6°C) for JBL3 by 2100 CE), which were
514 substantially warmer than either Joey Lake (mean annual temperature +1.5°C, range 0-3°C by 2100 CE)
515 or at Selwyn Lake (+0.5°C, Figure 4), were projected to lead to substantially more decomposition when
516 persisting for decades [e.g. *Schädel et al.*, 2016]. This is reflected in the magnitude of both net C losses
517 (Figure 5) and net C losses with depth at JBL3 (Figure 6A), which were much larger at this southern,
518 permafrost-free site compared to the cooler sites. The presence of permafrost in peat, rather than
519 enhance C loss with thaw, may instead have some potential to buffer peat C from decomposition in
520 warming temperatures through the heat sink of phase change in thawing ice in permafrost peat and the
521 underlying mineral soil and bedrock (e.g. Figure 4). This study showed smaller C losses from permafrost
522 peats relative to peats without permafrost (Figure 5) despite larger air temperature increases due to phase
523 change in frozen peat (Figure 4), which has also been shown recently more broadly across northern
524 peatlands [*Chaudhary et al.*, 2020].

525

526 *Key drivers of net peat carbon loss in observations and modeling: new peat additions vs. deeper peat*
527 *losses*

528 HPM-Arctic shows the importance of considering both C losses through enhanced
529 decomposition and the role of new peat accumulation at sites. The comparison of the two water table
530 scenarios, perched and flooded, at Joey, Selwyn, and Ennadai Lake demonstrate why net C losses are so
531 difficult to predict based on simple drivers: net C additions plays an important role in offsetting

532 decomposition losses, particularly in dynamic surface peats. Figure 6 shows the total old C losses
533 throughout the peat profiles (orange), while new peat additions occur mainly in the surface peat. For
534 Joey, Selwyn, and Ennadai Lake, the loss profiles (orange) look relatively similar between the perched
535 and flooded water table scenarios (Figure 6A, 5B, right and left panels), while net C added (blue) is
536 slightly larger in the flooded scenario across all sites (Figure 6, Figure 5). The net effect is a small
537 reduction in the net C loss in the flooded scenario relative to the perched scenario (Figure 5). Studies
538 based on measurements of net primary productivity and recent C accumulation rates in surface peat also
539 showed a net increase in C storage in thawed permafrost peats relative to permafrost peat plateaus for
540 western Canadian sites due to increased moisture and moss productivity [*Camill et al., 2001; Turetsky et*
541 *al., 2007*], which was simulated by the model (Figure S5). However, little is known about how net
542 primary productivity in peatlands changes in response to warming (and CO₂ fertilization) in either the
543 short-term or long-term [*Frolking et al., 2011*], introducing significant uncertainty into predictions of
544 change in net C stocks with permafrost thaw across the pan-Arctic [*McGuire et al., 2018*]. Recent
545 experimental show that new C inputs do not compensate for enhanced decomposition losses due to peat
546 warming [*Hanson et al., 2020*].

547 These modeling results illustrate the difficulties in trying to quantify losses of permafrost C from
548 peatlands using observations. HPM-Arctic shows that generally, net C additions occur in the surface
549 peat to 20 cm ($\sim +15 \text{ g C m}^{-2} \text{ y}^{-1}$), which could be observed using approaches to measure NPP and NEE
550 in surface peats (Figure 6). However, accounting for C losses in the deeper peat ($> 20 \text{ cm}$) changed the
551 direction of net C fluxes (Figure 6), given that predicted net losses of peat between 20 cm and 1 m were
552 substantial ($\sim -30 \text{ g C m}^{-2} \text{ y}^{-1}$); roughly 15% ($-5 \text{ g C m}^{-2} \text{ y}^{-1}$) of this could be attributed to permafrost C
553 loss (Figure 5). Approaches to quantify the response of peat C to changing environmental conditions
554 need to account for changes beneath the surface peat, particularly in the near surface peat (20-50 cm).

555 Given the lag between the warmest annual temperatures and the warmest peat temperatures in deeper
556 peat, deep peat and permafrost C losses could be highest in the fall and winter [*Estop-Aragonés et al.*,
557 2018b; *Webb et al.*, 2016] and may be difficult to capture with observations both due to timing and to
558 the relatively small magnitude of net C losses compared with the large annual fluxes of GPP and ER.
559 Additionally, we show that the contribution from permafrost and deep peat to net C losses are relatively
560 small compared to the shallower peats (Figure 6), thus the signal could be easily swamped by ecosystem
561 respiration in shallower peats.

562

563 *Insights and opportunities in peatland modeling*

564 HPM-Arctic combines a detailed process-based understanding of peat decomposition, hydrology,
565 C dynamics, and now how these interact with soil temperature and permafrost dynamics to influence
566 peat C accumulation and loss. For determining potential C losses in peatlands, a process-based
567 simulation model provides an alternative framework to empirical models using the permafrost thaw
568 chronosequences. A process model can explore consequences of different thaw scenarios, e.g., wetter or
569 drier conditions (Figure 5, 6). The HPM-Arctic model can explicitly separate the fate of carbon fixed
570 before or after a particular date, and so quantify loss of ‘old’ carbon and sequestration of ‘new’ carbon,
571 providing additional insights into the carbon dynamics (e.g. Figure 6). Process-based models are of
572 course limited by how well they represent the processes included, as well as any relevant processes that
573 have not been included. For example, the higher surface OM density at the northern sites is not
574 simulated by HPM-Arctic (Fig. S2), indicating that, in permafrost zones, there are likely processes in
575 peat accumulation and/or physical processing that are not included in the model or other models.
576 Another important dynamic that is not accounted for in this study is ice melt, a process not simulated in
577 HPM-Arctic, which can cause a range of effects, from peat compaction due to porewater ice melt that

578 can reduce active layer thickness [e.g. *Plaza et al.*, 2019] to abrupt, deep thaw in peatlands associated
579 either with melting of ice in the peat profile or in the underlying mineral soil. This abrupt thaw processes
580 occur in ice-rich permafrost deposits and can change deep peat temperatures and phase change from
581 permafrost to liquid water in a matter of years, making ‘old’ carbon more vulnerable to loss more
582 quickly, and has been a significant driver of C loss in other model simulations [e.g. *Nitzbon et al.*, 2020;
583 *Schneider von Deimling et al.*, 2015; *Turetsky et al.*, 2020; *Walter Anthony et al.*, 2018].

584 This study shows that net C balance in peatlands in response to warming can result in a range of
585 outcomes, from increased net C accumulation to net C losses. Across all sites, net C losses from active
586 layer peat between 0.2 – 1.0 m depth were the predominant driver of total C loss between 2015-2100
587 CE, rather than losses from deeper peat (>1 m) or newly thawed permafrost. Across the sites, new peat
588 accumulation offset a relatively large fraction of C losses, but the response of net primary productivity
589 to temperature for many peatland species and plant functional types remains poorly understood and is an
590 important area for future research. Factors such as site history, the presence/absence of permafrost in
591 the present-day, and climatic factors are important to consider when trying to predict how C stocks will
592 change with climate change in northern peatlands.

593

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